



Methylene Diphenyl Diisocyanate Reference Exposure Levels

Technical Support Document for the Derivation
of Noncancer Reference Exposure Levels
Appendix D1

**Public Review Draft Report
July 2014**



**Office of Environmental Health Hazard Assessment
California Environmental Protection Agency**

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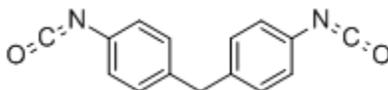
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Methylene Diphenyl Diisocyanate

Reference Exposure Levels

(Diphenylmethane diisocyanate, Methylene bisphenyl diisocyanate, 4,4'-Methylenediphenyl diisocyanate, Diphenylmethane-4,4'-diisocyanate)

CAS: 101-68-8



1. Summary

The Office of Environmental Health Hazard Assessment (OEHHA) is required to develop guidelines for conducting health risk assessments under the Air Toxics Hot Spots Program (Health and Safety Code Section 44360 (b) (2)). In response to this statutory requirement, OEHHA developed a Technical Support Document (TSD) that was adopted in 2008 and describes acute, 8 hour and chronic Reference Exposure Levels (RELs). The TSD presents methodology for deriving Reference Exposure Levels. In particular, the methodology explicitly considers possible differential effects on the health of infants, children and other sensitive subpopulations, in accordance with the mandate of the Children's Environmental Health Protection Act (Senate Bill 25, Escutia, chapter 731, statutes of 1999, Health and Safety Code Sections 39669.5 *et seq.*). These guidelines have been used to develop the following RELs for methylene diphenyl diisocyanate; this document will be added to Appendix D of the TSD.

Exposure to diisocyanates, including monomeric methylene diphenyl diisocyanate (MDI) and polymeric MDI, has been found to cause adverse effects on the respiratory system in both animals and humans. These effects include acute impacts such as sensory irritation and the induction of asthma in sensitive subjects. Chronic effects include long-term decrements in lung function and sensitization to asthma resulting in triggering of asthmatic attacks following very low exposures to diisocyanates (≤ 1 ppb). Literature summarized and referenced in this document covers the relevant published literature for MDI through Spring 2014.

1.1 Methylene diphenyl diisocyanate Acute REL

<i>Reference exposure level</i>	8 $\mu\text{g}/\text{m}^3$ (0.8 ppb)
<i>Critical effect(s)</i>	Increased total protein in bronchoalveolar lavage fluid of rats - marker of pulmonary irritation/injury
<i>Hazard index target(s)</i>	Respiratory system

1.2 Methylene diphenyl diisocyanate 8-hour REL

<i>Reference exposure level</i>	0.16 µg/m ³ (0.015 ppb)
<i>Critical effect(s)</i>	Bronchiolo-alveolar hyperplasia and pulmonary interstitial fibrosis
<i>Hazard index target(s)</i>	Respiratory system

1.3 Methylene diphenyl diisocyanate Chronic REL

<i>Reference exposure level</i>	0.076 µg/m ³ (0.0074 ppb)
<i>Critical effect(s)</i>	Pulmonary interstitial fibrosis
<i>Hazard index target(s)</i>	Respiratory system

2. Physical & Chemical Properties

Sources: HSDB (2013); US EPA, (1998b); Booth et al., (2009)

Chemical form	CAS	Vapor pressure
Methylene diphenyl diisocyanate monomer (MDI)	101-68-8	5×10 ⁻⁶ mm Hg @ 25°C, or 6.7×10 ⁻⁴ Pa @ 25°C
Polymeric methylene diphenyl diisocyanate (PMDI)	9016-87-9	2×10 ⁻⁶ mm Hg @ 20°C, or 3.1×10 ⁻⁴ Pa @ 20°C

<i>Description</i>	MDI: White waxy solid @ 20°C PMDI: Viscous amber- to dark-colored liquid @ 20°C
<i>Molecular formula</i>	C ₁₅ H ₁₀ N ₂ O ₂ (MDI)
<i>Molecular weight</i>	250.25 g/mol (MDI)
<i>Density</i>	1.23 g/cm ³ @ 25°C (MDI)
<i>Boiling point</i>	314°C (MDI)
<i>Melting point</i>	37°C (MDI)
<i>Saturated vapor conc.</i>	MDI: 60 µg/m ³ (6 ppb) @ 20°C PMDI: 32 µg/m ³ (3 ppb) @ 20°C
<i>Odor threshold</i>	odorless
<i>Solubility</i>	Soluble in acetone, benzene, kerosene, and nitrobenzene. Water solubility estimated at 1.51 mg/L at 25° C (MDI)
<i>Conversion factor</i>	10.24 mg/m ³ = 1 ppm @ 25° C (MDI)

3. Major Uses and Sources

Methylene diphenyl diisocyanate (MDI) is used in the preparation of polyurethane resin and spandex fibers, and to bond rubber to rayon and nylon. Its use in polyurethane foams accounts for approximately 80 percent of the MDI consumed

worldwide. The commercial form of MDI that is used in foaming operations is called polymeric MDI, or PMDI, and is typically a mixture of about 50 percent monomeric MDI and 50 percent higher molecular weight oligomers of MDI, mainly MDI dimers and trimers (Figure 1) (U. S. EPA, 1998b). MDI was introduced in the 1960s because it has a lower vapor pressure than toluene diisocyanate (TDI), which generated lower air concentrations of MDI relative to TDI during flexible foam operations. About 95 percent of all polyurethanes are based on MDI and TDI, with MDI used for the production of rigid polyurethane items (Vandenplas et al., 1993b). TDI is also listed as Toxic Air Contaminant and is described in a separate REL document.

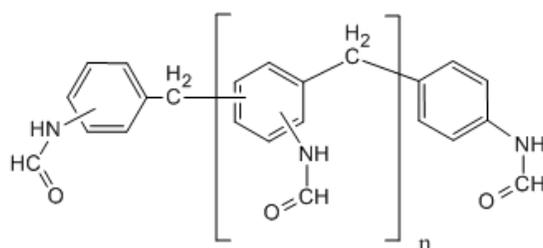


Figure 1 Structure of Polymeric MDI

Occupational exposure most commonly occurs during processes or applications in which the chemical is sprayed (mainly as an aerosol) or heated. With a vapor pressure of 5.0×10^{-6} mm Hg at 25°C, MDI will exist in both the vapor and particulate phases in the ambient atmosphere. Polyurethane spraying processes include spray-on truck bed lining, building insulation with sprayed-in-place polyurethane foam, and foam injection (Crespo and Galan, 1999; Ulvestad et al., 1999; Lofgren et al., 2003; Bonauto et al., 2005). MDI is also used in particle board bonding and production of mold cores in the foundry industry (Liss et al., 1988; Woellner et al., 1997).

Most studies that have collected personal breathing zone samples in the polyurethane foam industry have measured very low (often $<1 \mu\text{g}/\text{m}^3$) to non-detectable levels of MDI (Liljelind et al., 2010). In a study of a large body of industry air sampling data (8134 samples), most (74.6 percent) of the airborne MDI concentrations measured were below the limit of quantitation (LOQ) (Booth et al., 2009). Depending on the quantitation method, the LOQ was 0.04 to 0.5 $\mu\text{g}/\text{sample}$. However, only the monomer is typically quantified during air monitoring. Even though the higher molecular weight oligomers are also airborne during spray operations, analytical standards are not available to most laboratories to quantify these oligomers. In spraying operations, aerosolized airborne concentrations of MDI in the 0.077-0.400 mg/m^3 range have been measured (Crespo and Galan, 1999).

Vapor pressure studies using simultaneous torsion and mass loss effusion techniques showed that the molecular mass of the vapor phase above PMDI at 110°C was 250 ($\pm 7\%$), the same molecular weight as that of monomeric MDI

(Tury et al., 2003). This finding suggests that vapor released from heating processes in the manufacture of polyurethane from PMDI would consist primarily of monomeric MDI. While the size of the oligomer may affect its deposition and distribution in the lungs, a review of studies of MDI and PMDI suggests that at least the pulmonary effects of the two forms are expected to be qualitatively similar (Feron et al., 2001).

Vapor-phase MDI may be degraded in the atmosphere by reaction with photochemically produced hydroxyl radicals with an estimated half-life of 15 hours (Tury et al., 2003). Particulate-phase MDI is removed from the atmosphere by both wet and dry deposition. Mainly by analogy with studies on TDI, MDI is not expected to react significantly with atmospheric water vapor. When added to water (e.g., environmental spill), PMDI does not readily disperse and reacts slowly due to its high viscosity and low water solubility to form insoluble polyureas and only a very small amount of methylenedianiline (Yakabe et al., 1999). When added and mixed in water at very low concentrations (≤ 1 mg/L), higher conversions to diamines may be found. However, any diamines produced would be at relatively low concentrations, are aerobically biodegradable, and bind strongly and irreversibly to soil (Tury et al., 2003).

While occupational exposure occurs through inhalation of vapors and aerosols, and through dermal contact with compounds containing MDI, the general population may be exposed during the use of MDI-containing urethane adhesives, coatings, and foam products. Exposure to particulate and/or vapor phase MDI may also result from thermal decomposition of MDI-containing polyurethane foam as may occur, for example, during manufacturing, structural fires, or welding of polyurethane insulated pipe. Estimated facility emissions of MDI to the atmosphere in California was 3.6 tons per year in 2008, and 0.6 tons per year in a 2010 draft report (CARB, 2013).

4. Metabolism

Isocyanates, including MDI, are characterized by the $N=C=O$ group which contains two double bonds and exhibits strong chemical reactivity (Raulf-Heimsoth and Baur, 1998). Given its high chemical reactivity, inhaled MDI is expected to react initially with glutathione prior to being absorbed as the glutathione conjugate. Alternatively, a portion of the inhaled MDI may be cleared from the lungs and swallowed. If swallowed, conditions in the gastrointestinal tract favor spontaneous formation of polyureas, the smaller of which may be absorbed and excreted in the bile, while the larger urea polymers remain in the intestinal tract to be eliminated with the feces. The enzyme-catalyzed pathway of the proposed metabolic scheme (Figure 2) is expected to occur in the lungs, liver and/or kidneys following absorption and systemic distribution of MDI (Gledhill et al., 2005) with N-acetylation occurring prior to the hydroxylation step. The metabolic pathway shown in Figure 2 features monomeric MDI. It is not clear

how and to what extent the metabolism of the polymeric and monomeric forms may be different.

In a biomonitoring study of patients undergoing inhalation challenge tests with isocyanates, urinary MDI metabolites were collected and quantified following acid hydrolysis of the urine samples to form diphenylmethane diamines (Budnik et al., 2011). The urinary excretion peak of the MDI metabolite 4,4'-diphenylmethane diamine occurred 12-14 hrs after end of exposure. The urinary elimination of this metabolite was significantly slower than for other isocyanate amines, and excretion of the metabolite was not complete after 24 hrs. In another biomonitoring study, MDI metabolites in urine were found to reflect recent MDI exposure in workers during the past few days (Skarping et al., 1996). However, MDI metabolites in plasma reflected several weeks of exposure, likely a result of isocyanate adduct formation with blood proteins.

A study of genotypic variation in enzymes involved in the metabolism of MDI, specifically N-acetyltransferases (NATs) and glutathione transferases (GSTs), among occupationally exposed workers revealed a complex picture (Littorin et al., 2008). For example, two of four GST genotypes, GSTP1¹¹⁴ and GSTP1¹⁰⁵, were associated with higher levels of urinary metabolites of MDI than were the other two. At the same time, GST1¹⁰⁵ was associated with lower levels of serum MDI-specific IgG and fewer eye symptoms, but with an increased risk of symptoms in the airways, as well as with atopy. The allergic symptomatology appears to be affected by how rapidly MDI is conjugated to glutathione for excretion. By comparison, among workers with slow NAT2 acetylating capacity, lower plasma and urinary levels of MDI metabolites, lower MDI-specific IgG levels, and better lung function were observed, but with a higher risk of airway and eye symptoms. Thus the variations among workers in the manifestation of pulmonary and allergic symptoms following MDI exposure reflect the complex genotypic variation in metabolic enzymes, and the speed with which MDI is removed from the system.

As described above, MDI reacts with GSH in lung lining fluid that can be then absorbed into the bloodstream. *In vitro* studies have shown that GSH can act as a "shuttle" for MDI, in that once MDI-GSH is absorbed MDI-albumin conjugates are generated via GSH-mediated transcarbamoylation, which exhibit distinct changes in conformation and charge (Wisnewski et al., 2013). These MDI-albumin conjugates were specifically recognized by serum IgG of MDI workers with diisocyanate-induced asthma, suggesting one possible pathway for MDI in promoting immune responses.

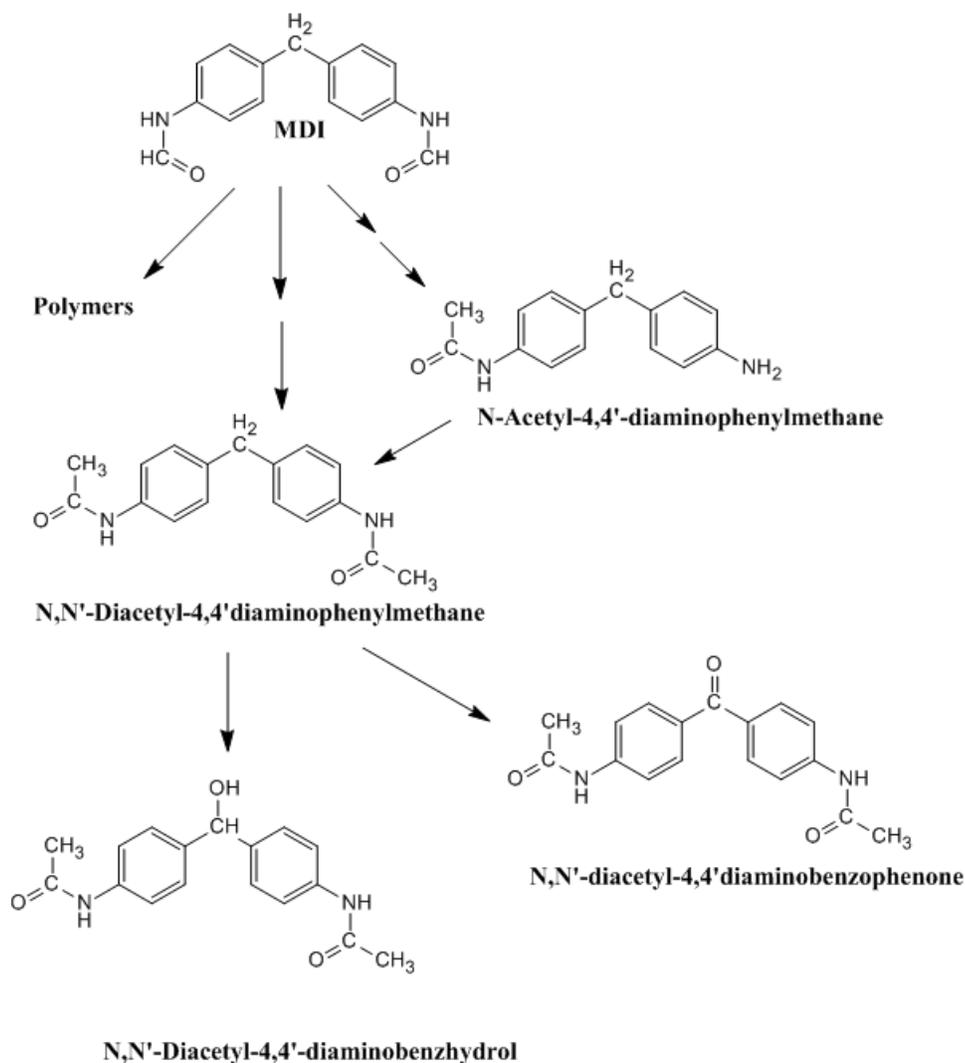


Figure 2 Metabolic Scheme for Monomeric MDI

5. Acute Toxicity of Methylene Diphenyl Diisocyanate

As is the case with other diisocyanates such as toluene diisocyanate (TDI), MDI has the capacity to cause sensitization of the neuroimmune system. Several isocyanates and other reactive electrophiles have been shown to activate cation channels of the transient receptor potential A (TRPA) group in sensory neurons (Macpherson et al., 2007; Taylor-Clark et al., 2009). This can lead to long-term potentiation of synapses in the brainstem, and subsequent airway hyperresponsiveness. In addition, neuropeptides released during MDI stimulation of sensory neurons may cause mast cell degranulation, goblet cell hyperplasia and mucus secretion, contraction of airway smooth muscles, and pulmonary edema. However, MDI is less potent than TDI in causing these

effects. Diisocyanates activate TRPA channels in nociceptive neurons in the airways leading to respiratory symptoms via long-term potentiation of neural pathways, release of inflammatory mediators, and stimulation of the immune system.

Asthmatic cross reactivity between different isocyanates has been documented. Innocenti et al. (1988) found that nearly 50 percent of subjects with asthma induced by TDI also exhibited asthmatic reactions to MDI, to which they were never exposed at work. In another study, of 13 workers exclusively exposed to MDI, four also reacted to TDI (O'Brien et al., 1979). In six workers with IgE-mediated sensitization to isocyanates, radioallergosorbant test (RAST) and/or skin test investigations revealed the presence of IgE antibodies reacting specifically with human serum albumin (HSA) conjugated with those isocyanates to which workers were exposed as well as with other isocyanates with which they had not been in contact (Baur, 1983). These results indicate the predominance of closely related antigenic determinants in HSA conjugated with different isocyanates. The common antibody-binding regions are recognized to different extents by antibodies of clinically sensitized workers, indicating individual differences in specificities and avidities of antibody populations.

5.1 Acute Toxicity to Adult Humans

Acute inhalation exposure to MDI generally results in irritation of the lungs and upper respiratory tract with symptoms including headache, sore throat, cough, and chest tightness.

Four specific types of respiratory health ailments resulting from worker exposure to diisocyanates have been described in Latza et al. (2002):

- Occupational asthma without a latency period (RADS)
- Occupational asthma with a latency period
- Hypersensitivity pneumonitis or extrinsic allergic alveolitis
- Chronic obstructive lung disease

If the initial acute exposure is high enough, a nonimmunological type of asthma may occur encompassing irritant-induced asthma or reactive airways dysfunction syndrome (RADS). Subsequent low-level MDI exposures in these individuals result in pulmonary symptoms including bronchial hyperresponsiveness and airflow obstruction (Leroyer et al., 1998). Occupational asthma with a latency period and hypersensitivity pneumonitis generally occur with chronic or multiple exposures to MDI and other diisocyanates. Once sensitized, the individual may experience these symptoms with acute low-level exposure to MDI.

A case report that is illustrative of an acute high exposure resulting in RADS is that of a foundry worker who had frequent exposure to MDI but no reported respiratory or other symptoms. After three years he received an intense acute

inhalation exposure as a result of an MDI spill in his work area. Within one hour he experienced headache, sore throat, cough and chest tightness. Other workers in the area experienced similar symptoms but only transiently. These initial symptoms were consistent with a diagnosis of RADS. However, over the course of the subsequent month his chest symptoms and wheezing worsened, especially at work, with some remission during weekends. Spirometric testing revealed moderate airflow obstruction, a forced expiratory volume in 1 second (FEV₁) of 2.5 l (83% predicted), and a forced vital capacity (FVC) of 4.5 l (121% predicted). Symptoms persisted despite treatment with budesonide, a glucocorticoid anti-inflammatory. After salbutamol inhalation (a β_2 -adrenergic receptor agonist to treat bronchospasm), FEV₁ increased 12 percent. Bronchoprovocation tests with 15 ppb MDI were performed for 4, 30 and 60 min. An isolated late reaction was associated with the 60 min exposure, with a 22 percent fall in FEV₁ seven hours after exposure. The authors suggest his symptoms were consistent with occupational asthma caused by the acute high level exposure (Leroyer et al., 1998). However, it is not clear what role the low level exposures prior to the acute exposure may have played in the etiology of this case's symptomatology.

Provocation tests have been used to confirm a diagnosis of MDI-induced occupational asthma in polyurethane workers. However control groups consisting of normal or asthmatic individuals without previous exposure to isocyanates were not included in these studies to quantitatively elucidate potential acute toxicity of MDI exposure. In the provocation studies, concentrations of MDI or PMDI used in challenge tests ranged between 1 and 20 ppb with exposure durations of seconds (i.e., one breath) to up to 4 hrs (O'Brien et al., 1979; Burge, 1982; Zammit-Tabona et al., 1983; Cartier et al., 1989; Vandenplas et al., 1992; Leroyer et al., 1998; Piirila et al., 2000; Lemiere et al., 2002).

These studies have observed asthmatic responses to exposures of 1 ppb MDI or lower in sensitized workers undergoing challenge testing. Lemiere et al. (2002) exposed eight subjects with occupational asthma induced by specific diisocyanates to 1 ppb MDI, TDI or hexamethylene diisocyanate (HDI) using a closed circuit apparatus. The authors considered a positive result to be a 20% or greater reduction in FEV₁. By this criterion asthma was induced in two of the subjects with a 30 min exposure, one to MDI and the other to HDI. A third subject had asthma induced with a 45 min exposure to TDI. There was also a significant correlation (Spearman rank order test $p=0.8$, $P<0.001$) between the percentage of maximum decrease in FEV₁ after exposure to 1 ppb and the increase in sputum neutrophil count, indicating inflammatory changes as well. In another study, Burge (1982) found that 2 of 24 MDI-sensitized workers showed a positive reaction with exposure to MDI as low as 1 ppb. The criterion in this study for a positive reaction was a 15 percent or greater reduction in FEV₁.

The lowest concentration of MDI resulting in an asthmatic response in a sensitized worker occurred following a 15 min exposure to 0.51 $\mu\text{g}/\text{m}^3$ (0.05 ppb) monomeric MDI (Suojalehto et al., 2011). The subject's FEV₁ fell by a maximum of 25% from base line after 1 hr, requiring use of a bronchodilator and an oral steroid. The worker had a history of severe reactions during work, in which she was occasionally exposed to synthetic plastic containing MDI during orthopedic plaster casting. The levels of MDI in the air were measured in the exposure chamber using filter collection and subsequent liquid chromatography-mass spectrometry analysis of the MDI isocyanate groups. Sampling of breathing zones of nurses when applying and removing the casts was below 1 ppb MDI. The authors noted that dermal exposure to the unhardened MDI-containing plastic material could have been a factor in development of respiratory sensitization.

The study by Suojalehto et al. (2011) provides evidence that it is likely not possible to set a REL that can protect every individual that has acquired specific hypersensitivity to MDI, due to some sensitized individuals having a positive response at extremely low concentrations.

In a study by Zammit-Tabona et al. (1983), exposure to an average MDI concentration of 12 ppb over 60 min resulted in a ≥ 20 percent fall in FEV₁ in 7 of 11 foundry workers suspected to have MDI-induced asthma. One of the responders developed cough and wheeze after 10 min of exposure to MDI but recovered within 5 min after leaving the chamber. The MDI concentration in the chamber had only reached 10 ppb when the subject had a positive response. This subject also responded in a similar fashion to 2.5 ppm formaldehyde after 10 min of exposure, and recovered within 5 min of cessation of exposure. None of the other 10 subjects responded with an asthmatic reaction to formaldehyde exposure. This subject had a longstanding history of asthma before employment at the foundry and had a marked degree of bronchial hyperreactivity to methacholine. The authors concluded the cause of bronchoconstriction in this subject from both MDI and formaldehyde was likely irritation and not sensitization. This study suggests that MDI can induce an asthmatic response in individuals with asthma due to MDI's pulmonary irritant qualities, rather than sensitization from previous exposure to diisocyanates.

5.2 Acute Toxicity to Infants and Children

Asthma-like symptoms were observed among 203 school children following acute exposure to MDI spilled during track paving with polyurethane (Jan et al., 2008). Of the exposed children, 70.9 percent reported headache, 67.5 percent had persistent cough, 63.5 percent had dyspnea, and 62.6 percent had nausea. Chest discomfort was reported by 23.6 percent of the students but chest X-rays were normal. Bronchodilators were administered to 15.8 percent who experienced wheezing and difficulty breathing. Although no measurements of

actual air MDI levels were reported, the authors observed an inverse linear relationship between the incidence of affected students in various classrooms and the distance from the site of MDI spillage ($r = -0.48$, $p < 0.05$) suggesting a dose-response.

During follow-up surveillance three days after the incident, the prevalence of residual symptoms was cough 30.0 percent, headache 19.7 percent, dyspnea 15.3 percent, sore throat 10.3 percent, and nausea 3.9 percent (Jan et al., 2008). A positive history of asthma among 10.8 percent of the students was strongly correlated with the incidence of dyspnea (OR 4.09; 95% CI 1.17-14.32) and an abnormal pulmonary function test (OR 3.84; 95% CI 1.09-13.5). However, none of the other symptoms during the episode was correlated with either asthma history or abnormal lung function tests. In addition, 60.8 percent of the children without a history of asthma also complained of dyspnea, and 16.2 percent required bronchodilators for symptomatic relief. Acute exposure to high levels of MDI was thus associated with an asthma-like syndrome among previously unexposed individuals. The authors did not discuss effects seen in exposed adults, so it is unclear if children were more prone to the acute effects of MDI than adults.

Krone and associates have postulated that a relationship exists between exposure to polyurethane products made from isocyanates and childhood asthma (Krone and Klingner, 2005). Further discussion is presented in Section 6.2.

One animal study was found that investigated the differential sensitivity of young rats to PMDI. In a subacute study, four week old and six week old rats (20 rats/group/sex) were exposed to 14.1 mg/m^3 PMDI for 6 hr/day, five days/week for 2 weeks (Reuzel et al., 1994b). Only mortality was recorded. Four-week-old rats died earlier and in greater numbers than did rats that were six weeks old. This early-life susceptibility was greater for females than for males. The reason for differential sensitivity was unknown to the authors. However, in humans the relative minute volume to surface area in the pulmonary region is greater in infants by 3-4-fold compared to adults (OEHHA, 2008). Chemicals that are pulmonary irritants, such as MDI, are predicted to have greater pulmonary effects in the young.

5.3 Acute Toxicity to Experimental Animals

During a single 4 hour exposure to high concentrations of PMDI aerosols ($376 - 638 \text{ mg/m}^3$, particle size $< 5 \mu\text{m}$), Wistar rats displayed labored respiration and mouth breathing (Reuzel et al., 1994b). Deaths occurred at all exposure levels within two days following the end of exposure, with an LC_{50} of 490 mg/m^3 . Among the animals that survived, transient weight loss was observed during the second and fourth days after exposure. At the higher aerosol levels, the lungs of

rats euthanized immediately after exposure were grayish and wet, with some pulmonary hemorrhaging and hemorrhagic nasal discharge.

A nose-only exposure 4-hr lethal potency comparison study of PMDI and monomeric MDI was carried out by Pauluhn (2011). The LC₅₀ of rats exposed to PMDI (mean particle size 1.8 µm) was 310 mg/m³, with males and females equally sensitive. The LC₅₀ of MDI (mean particle size 3.3 µm) was 367 mg/m³ (males and females combined), with males approximately twice as susceptible as females.

As with the acute exposures above, subacute exposure to PMDI aerosols (0, 2.2, 4.9, 13.6 mg/m³; 10 rats/sex/dose) for six hr/day, five days per week for two weeks led to labored respiration and mouth breathing in the high exposure group (13.6 mg/m³) starting on day four (Reuzel et al., 1994b). Other clinical changes reported for this group included slow movements, dyspnea, piloerection, salivation, bleeding from the nares and swollen abdomens. While rats in the 2.2 mg/m³ group were not visibly affected, those in the 4.9 mg/m³ group were restless, slightly dyspneic, and showed piloerection. In all treatment groups, lung weights were elevated relative to body weights. The effects of PMDI were mainly on the respiratory tract with males more severely affected than females.

The adverse effects of acute exposures to PMDI manifest mainly in the lungs as pulmonary inflammation characterized by increased immune cell infiltration, protein production, and organ weight. Kilgour et al. (2002) examined the appearance and resolution of these effects over a 30 day period in rats following an acute 6 hour exposure to 10, 30, or 100 mg/m³ PMDI. Immediately following a single, 6-hr acute exposure, lung lavage fluid showed massive increases in neutrophils (37 percent of total cells at 10 mg/m³, 78 percent at 100 mg/m³), but reduced numbers of alveolar macrophages. Protein content was elevated in lavage fluid, and enzyme activities increased for N-acetyl glucosaminidase (NAG), alkaline phosphatase, and lactate dehydrogenase (LDH). The accumulation of crystalline surfactant and cellular debris in alveolar lumina through day three at all PMDI concentrations suggests PMDI is cytotoxic to macrophages. By day three post-exposure, neutrophil numbers were still markedly elevated in the 100 mg/m³ group, but had dropped substantially in the lower dose groups, while macrophage numbers increased. Lactate dehydrogenase continued to rise but the NAG and alkaline phosphatase activities had returned to control levels. By day ten, most of the measured parameters had returned to control levels, although epithelialization of the alveoli was observed in animals at 30 and 100 mg/m³. Thirty days following the last exposure, lung weights, lung lavage parameters, cell proliferation and ultrastructural appearance had returned to normal. These results suggest that even at relatively high acute exposure levels, recovery from PMDI-associated toxic effects in the lung is relatively rapid.

Many of the effects observed by Kilgour et al. (2002) may be related to MDI-induced changes in the pulmonary epithelium that forms the blood-air barrier in the lungs. Pauluhn (2000) examined bronchoalveolar lavage fluid (BALF) of female Wistar rats (n=6 or 7) for markers of damage to pulmonary epithelium following an acute 6-hr exposure to MDI at 0.7, 2.4, 8, or 20 mg/m³. These markers included angiotensin converting enzyme (ACE), protein levels, alkaline phosphatase, lactate dehydrogenase, γ -glutamyltranspeptidase, and sialic acid, and were assayed 3 hours, 1, 3, and 7 days following exposure. PMDI at all dose levels caused an immediate significant increase in alkaline phosphatase activity ($p < 0.05$) that returned to control levels by day three. This was deemed to be consistent with an adaptive increase in pulmonary surfactant that is rich in alkaline phosphatase from type II pneumocytes. Plasma protein levels in BALF similarly were immediately and significantly elevated at all PMDI concentrations suggesting damage to the epithelial barrier. The activity of ACE was also significantly elevated but only at 2.4 mg/m³ and above. However, there was not a dose-dependent effect on LDH levels, suggesting that cytotoxicity was not the cause of the elevated protein and ACE levels. Glutathione levels measured in BALF peaked on day one following exposure, and returned to control levels by day seven. However, GSH levels in lung tissue remained elevated through day seven.

Based on these results, Pauluhn (2000) proposed that MDI interferes with the pulmonary surfactant system leading to surfactant dysfunction and increased alveolar surface tension. This surface tension in turn enhances transudation of fluid and solutes from the capillaries, contributing to the pulmonary edema characteristic of MDI exposure. While this effect appears to be transitory at these dose levels, it was observed to occur at concentrations as low as 0.7 mg/m³.

Pauluhn (2002) conducted a concentration \times time ($C \times t$) study in which male rats were exposed to 3.4 to 58.1 mg PMDI/m³ and exposure durations of 6 hr to 23 min, respectively, so that $C \times t$ was approximately 1200 mg/m³-min. Using total protein content in BALF one day post-exposure as the endpoint for response, total protein was increased 50 percent for all PMDI exposure groups compared to the control group. Thus, the magnitude of BALF protein matched the exposure intensity over the entire range of concentrations investigated when $C \times t$ was kept constant. The conclusion was that changes in BALF protein were dose-dependent (i.e., equally dependent on changes in concentration and duration of exposure).

In a separate study, groups of female Wistar rats were exposed to a range of PMDI concentrations (0.7, 2.3, 8 and 20 mg/m³) for 6 hrs (Pauluhn, 2000; 2002). Earlier results showed total protein and ACE in BALF were among the most sensitive indicators for acute irritant effects of PMDI. Total protein and ACE in BALF were determined at 3 hrs and 1 day post-exposure. Total protein at 1 day post-exposure and ACE at 3 hr and 1 day post exposure were statistically

significantly increased above control following exposure to 2.3 mg/m³ PMDI (Table 1). Total protein 3 hr post-exposure was statistically significantly increased above control at the lowest exposure of 0.7 mg/m³ PMDI. At 1 day post-exposure, total protein in BALF had returned to control level in rats exposed to 0.7 mg/m³ PMDI, while total protein levels were still increased and essentially unchanged in rats exposed to higher PMDI concentrations. Double-logarithmic analysis of the concentration-effect relationship for these two endpoints estimated an acute irritant benchmark no-effect threshold concentration of 0.5 mg/m³ for 6 hr exposure to PMDI aerosol. The no-effect threshold was considered a relative change of 100 percent total protein and ACE from control values.

We applied continuous modeling methodology in U.S EPA's benchmark dose software, version 2.3.1 (U. S. EPA, 2012) for benchmark dose analysis on the statistical results for increased total protein and ACE in BALF with increasing PMDI exposure. The data in Table 1 was kindly provided to OEHHA by Dr. Pauluhn in order to run BMD modeling and originates from Pauluhn (2002) in which the data had been presented only in graphical form.

Table 1. BALF results for total protein and ACE following 6 hour exposure of rats to PMDI

Endpoint	Dose (mg/m ³) ^a				
	0	0.7	2.3	8	20
Total protein – 3 hr post-exposure (g/l) mean±SD	0.152 ±0.034	0.224 ±0.021	0.215 ±0.037	0.363 ±0.062	0.484 ±0.131
Total protein – 1 d post-exposure (g/l) mean ± SD	0.153 ± 0.018	0.164 ±0.016	0.222 ±0.030	0.336 ±0.066	0.464 ±0.139
ACE – 3 hr post-exposure (nmol/min/ml) mean ± SD	0.099 ±0.047	0.106 ±0.038	0.209 ±0.017	0.369 ±0.069	0.411 ±0.202
ACE – 1 d post exposure (nmol/min/ml) mean ± SD	0.102 ±0.043	0.096 ±0.011	0.210 ±0.055	0.249 ±0.055	0.436 ±0.134

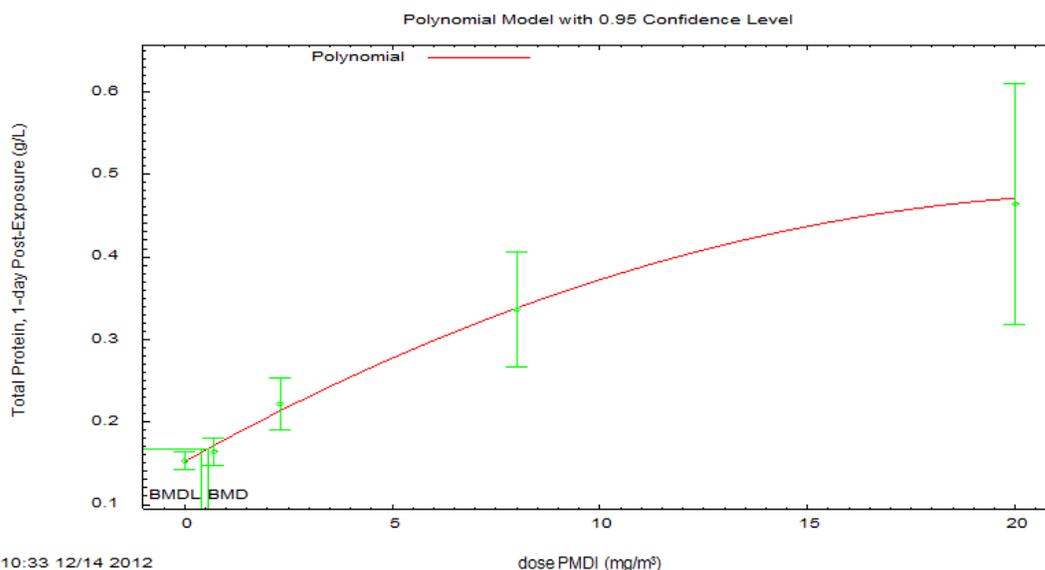
^a n per dose group: n=12 for 0 mg/m³; n=6 for all other exposure groups.

No model in the U.S. EPA BMD suite was able to fit an acceptable line to the data points for ACE 3 hrs and 1 day post-exposure, and total protein 3 hrs post-exposure. Although a statistically significant dose-response was found for these endpoints (p<0.05), the small n and variations in means and variances at the exposure levels resulted in unacceptable fits with the models provided. However, acceptable non-homogeneous variance model fits were found for total protein at 1 day post-exposure.

A one standard deviation (SD) from the control group was determined for the benchmark dose concentration (BMC) of total protein at 1 day post-exposure. The 95% lower confidence limit on the BMC (the BMCL) was used as the point of departure (POD). Following U.S. EPA guidelines, we chose the model with the lowest AIC (Akaike information criterion) in instances where various acceptable model fits to the data were similar. The BMCL chosen for the POD was the 2nd

degree polynomial model (At 1 SD: BMC= 0.5750 mg/m³, BMCL= 0.4068 mg/m³), which had the lowest AIC among the models that could be fit to the data. The model fit to the data is graphically illustrated in Figure 3.

Figure 3. Second degree polynomial model (1 SD for BMC) fit to the data for total protein in BALF 1 day following 6 hr exposure to 0.7 to 20 mg/m³ PMDI



The effects of 4 hr exposures to MDI aerosol (mass median aerodynamic diameter 0.7 μm and geometric SD 1.6 μm) on pulmonary and sensory irritation at concentrations ranging from 7 to 59 mg/m³ were measured as respiratory rate depression in mice (Weyel and Schaffer, 1985). The concentration required to reduce the respiratory rate 50 percent (RD50) was 32 mg/m³. Respiratory depression progressed slowly with MDI exposure, not reaching a plateau until 3-4 hrs into the exposures. The decline in respiratory rate was dependent on both concentration and duration. Unlike other isocyanates such as TDI and HDI, MDI acted primarily as a pulmonary irritant evoking little or no sensory irritation (i.e., stimulation of lower respiratory tract receptors and not the trigeminal nerves in the upper respiratory tract). Increased lung wet weight was observed at every concentration 24 hrs after the exposures.

Several animal models of asthma have been developed for both respiratory and dermal sensitization to MDI or PMDI (Ratray et al., 1994; Pauluhn et al., 2000; Pauluhn and Poole, 2011; Wisnewski et al., 2011). In guinea pigs, one high level 15 min exposure to 135 mg/m³ PMDI resulted in respiratory sensitization when challenged three weeks later with 15 and 49 mg/m³ PMDI (Pauluhn et al., 2000). Respiratory sensitization was assessed by a change in respiratory rate and an influx of eosinophilic granulocytes into bronchial tissues. Another study in guinea pigs using a lower dose of 19.4 to 23.7 mg/m³ MDI 3 hrs/day for 5 consecutive days did not lead to sensitization (Ratray et al., 1994).

In a Brown Norway rat asthma model, the $C \times t$ relationship for PMDI sensitization was examined using a 5-day exposure sensitization protocol (Pauluhn and Poole, 2011). Consistent with the sensitization protocol used, the most sensitive endpoints characterizing an allergic pulmonary inflammation were BALF-neutrophils and delayed-onset respiratory changes. A high exposure concentration of MDI during 10 min exposures (e.g., 100 mg/m³ for 10 min) elicited a more vigorous response than the similar $C \times t$ at 360 min (e.g., 3 mg/m³ for 360 min). The $C \times t$ study also showed that the dose that triggers an elicitation response in the rat asthma model is slightly below that causing acute pulmonary irritation in naïve rats. This finding suggests that allergic responses via inhalation of PMDI appear to be linked with pulmonary irritation of the lower airways.

Animal studies have shown that MDI skin exposure can induce MDI sensitivity with subsequent challenge via the respiratory tract, suggesting that skin contact may be an important cause of occupational respiratory asthma. In mice and guinea pigs with previous MDI skin exposure ($\geq 1\%$ MDI in solution) significant airway inflammatory responses to respiratory MDI challenge have been demonstrated (Rattray et al., 1994; Wisnewski et al., 2011). These inflammatory responses included influxes of eosinophils and lymphocytes in BALF samples and serum antibody responses.

6. Chronic Toxicity of Methylene Diphenyl Diisocyanate

6.1 Chronic Toxicity to Adult Humans

The effects of chronic exposure to MDI are largely reflected in decrements in pulmonary function and exacerbation of asthma. Impairment of lung function is mainly a function of allergic inflammation. Variable airflow restriction of the airways and bronchial hypersensitivity are associated with asthma.

On rare occasions isocyanates can also cause extrinsic alveolitis, or hypersensitivity pneumonitis, which involves the air sacs and lung parenchyma. Symptoms of hypersensitivity pneumonitis include headache, nausea, muscle aches, fever and chills, significant falls in both FEV₁ and FVC, hypoxia, audible moist rales, increased blood neutrophils, increased neutrophils and lymphocytes in bronchoalveolar lavage, and significant levels of IgG and IgE antibodies to MDI-human serum albumin (Baur et al., 1984; Vandenplas et al., 1993a).

MDI appears to have a greater propensity to cause hypersensitivity pneumonitis than TDI (Vandenplas et al., 1993a; Baur, 1995). At least 4.7 percent of workers exposed to PMDI resin in the manufacture of woodchip boards developed hypersensitivity pneumonitis (Vandenplas et al., 1993a). The latency period before symptoms appear ranged from weeks to 1 year following beginning of exposure.

It is not clear from occupational studies whether asthma and hypersensitivity pneumonitis are more the result of chronic low-level exposure, acute exposure to high levels, or both. It is clear, however, that once individuals are sensitized to MDI, further exposure generally exacerbates respiratory symptoms (Vandenplas et al., 1993b; Piirila et al., 2000).

A 10-year follow-up of 245 workers that had been diagnosed with asthma induced by diisocyanates, 96 of which were due to MDI exposure, found a generally poor medical outcome (Piirila et al., 2000). Some patients (15 percent) reported occasional isocyanate exposure in their current work. Of the patients 82 percent still experienced symptoms of asthma, 34 percent used no medication and 35 percent were on regular medication. However, FEV₁ reduction did not exceed the predicted decline over time in either smoking or nonsmoking patients.

Longitudinal studies are the primary means for assessing asthma onset prevalence and changes in pulmonary function with time in diisocyanate workers. There are numerous longitudinal studies examining pulmonary changes in TDI workers. However, there have been few longitudinal studies examining the effects of MDI in workers. In addition, the air concentrations of MDI in occupational settings are often poorly characterized. This is due, in part, to airborne MDI concentrations below the detection limit of the samplers. MDI also forms a considerable amount of dimers and polymers that are not adequately evaluated by routine measurements, but likely cause similar work-related symptoms as MDI (Baur et al., 1994). Consequently, MDI concentrations may be underestimated.

The following summaries represent the most comprehensive occupational studies available, most of which included limited MDI exposure data.

Pham et al. (1987)

Chronic exposure to mainly MDI was studied in a five-year longitudinal study of workers from two factories producing polyurethane foam. A respiratory questionnaire, measurement of vital capacity (VC) and FEV₁ and a single breath CO diffusion test (DL_{CO}) were done at the beginning of the survey, and then repeated five years later. DL_{CO} lung diffusion testing is a measure of how well the lungs exchange gases. A total of 318 workers participated, 83 of whom were unexposed, 117 indirectly exposed and 118 directly exposed to MDI. The MDI concentration was characterized as below 20 ppb. The authors did not state what proportion of diisocyanates was MDI; only that it was mainly MDI.

At the beginning of the study, both indirectly and directly exposed workers reported greater symptoms of chronic bronchitis compared to controls ($p < 0.05$), but only directly exposed female workers reported greater symptoms of asthma ($p < 0.05$). Pulmonary function tests of directly exposed men showed a lower

percent of predicted VC, FEV₁ and DL_{CO} (p<0.01). These pulmonary decrements were most pronounced in those workers with >60 months of direct exposure.

Five years later, only half of the initial cohort was still active (114 males and 45 females). Asthma and chronic bronchitis had increased in both exposed groups and the unexposed group. No difference in these symptoms was observed between groups. The five year decline in VC and FEV₁ was not significantly different between groups, but a significantly larger loss of DL_{CO} (p<0.05) was found in the directly exposed workers.

Petsonk et al. (2000)

In a prospective study of the respiratory effects of MDI exposure, Petsonk et al. (2000) evaluated the respiratory health of workers in a new wood products manufacturing plant in which MDI was used as a binder. Health data and exposure histories were collected by questionnaire prior to the use of MDI at the plant, and semiannually for the next two years. The critical effect was asthma-like symptoms, cases of which were defined based on questionnaire responses as current or previous asthma, or current use of a bronchodilator, or current asthma attacks characterized by shortness of breath and wheezing.

Cases were divided into those that met these criteria at the initial survey (IAS) before MDI exposure, those who met the definition during a follow-up survey (FAS) after exposure to MDI had begun, and a third group with new onset asthma-like symptoms (NAS), defined as workers who met the case definition for FAS, but not for IAS. Measurements of serial peak flow, spirometry, methacholine challenge, and specific IgE were used in some cases for validation of case designation, but were not available for all study participants. Thus, the authors noted that it was unlikely that all participants with respiratory symptoms have occupational asthma.

Of the 178 workers with initial and at least one follow-up survey, a complete occupational history was available for 144. Of these, 77 completed the initial health survey prior to first use of MDI at the plant and also reported no previous job with MDI exposure. Thus the remaining 67 may have had MDI exposure at the plant prior to their initial health assessment. Thirty-two workers (20%) met the FAS case definition, while 22 workers (12%) met the NAS case definition. In the NAS group, the duration of work prior to symptom onset ranged from 3 to 22 months (mean, 11 months). The prevalence of FAS and NAS cases was clearly associated with reported exposure in that those who reported working with MDI were significantly more likely to have asthma (p < 0.01) than were those with no or only occasional passing exposure. Those working in areas with high potential exposure to liquid MDI (e.g., cleanup of MDI spills and cleaning the MDI blender) had a significantly elevated prevalence of asthma (p < 0.001) compared to those where potential exposure was rated medium or low. Both FAS and NAS cases were significantly elevated among those who indicated they occasionally

removed protective respirators compared with those who never did ($p = 0.05$), and by 52 percent of those who reported at least once observing MDI stains on their skin ($p < 0.001$).

These observations in conjunction with the controls engineered into the plant's design to reduce inhalation exposure suggest that the appearance of new asthma symptoms among a third of those working in the blending and press operations, and 10-30 percent of the workers in adjacent areas, likely reflects both inhalation and dermal exposure to MDI. Environmental sampling for diisocyanate was not carried out during the study. Six personal breathing zone samplers (OSHA method 47) worn by employees 7 months after the last health survey did not find detectable air levels of MDI. One wipe glove sample did find 78 μg of MDI. The authors suggested the high proportion of asthma-like symptoms was more related to many participants working with or cleaning spills of liquid MDI, rather than to long-term exposure to low level airborne MDI.

Johnson et al. (1985)

A cross-sectional study was conducted in 78 iron and steel workers exposed to Pepset, a chemical binding agent consisting of MDI, phenol, formaldehyde and their decomposition products, and silica-containing particulates. A group of 372 railway yard workers matched for socioeconomic status and smoking habit, and "without significant exposure to air contaminants" (as determined by environmental measurements taken during the health study) were used as controls [OEHHA notes that railway yard workers are often exposed to diesel engine exhaust, and prior history of diesel exhaust exposure in this control group was not reported]. Exposure to MDI was carried out during the health survey portion of the study by area sampling ($n=319$) of multiple sites in the foundry with midget impingers. A colorimetric method was used for analysis of MDI air concentrations. Sampling times were not given. Of the area samples collected, 85.6 percent were <5 ppb, 8.5 percent were >5 ppb and ≤ 10 ppb, 4.4 percent were >10 and ≤ 15 ppb, 0.9 percent were >15 and ≤ 20 ppb, and 0.6 percent were >20 ppb (2 out of 319 samples). The authors noted that levels in excess of 20 ppb were more common before a new ventilation system was installed several months before the study.

For prevalence of respiratory symptoms, phlegm, breathlessness, chest tightness and chest illness were statistically significantly greater ($p < 0.05$) in foundry workers compared to the control group. The prevalence of wet cough was also significantly higher in foundry workers compared to controls. The mean FEV_1 , FVC and $\text{FEF}_{25-75\%}$ of foundry workers were all significantly lower than those of controls. As expected, current smokers had significantly worse mean lung function compared to non-smokers. The foundry workers also underwent a methacholine challenge test. A provocative concentration of ≤ 8 mg/ml causing a 20 percent or greater drop in FEV_1 ($\text{PC}_{20} \leq 8$ mg/ml) in a worker was chosen as an indicator of bronchial hyperreactivity. By this criterion, 19.7 percent (13 of 66

workers) had evidence of bronchial hyperreactivity. Three workers had evidence of pneumoconiosis in chest radiographs, suggestive of silicosis. The authors noted that exposure to dusts and other chemicals could have been a contributing factor to the reduced pulmonary function of the workers.

Inhalation provocation tests with MDI and formaldehyde were performed on nine of the asthmatic workers in a separate study (Zammit-Tabona et al., 1983). Six had a positive asthmatic reaction to MDI exposure, but not to formaldehyde, indicating MDI was the cause of their asthma. A seventh worker had an immediate reaction to both MDI and formaldehyde. But the transient bronchoconstriction due to exposure in this worker was likely due to irritation rather than sensitization.

Bernstein et al. (1993)

A cross-sectional study was conducted in 243 workers exposed to MDI in a urethane plant that had been designed to minimize MDI exposure. There were 147 workers on the urethane mold lines and 96 other workers were involved with administrative, transport, or maintenance activities. The average duration of employment in the plant was 18.2 months (range 0 to 32 months).

Exposure to MDI fumes was continuously monitored via a visible spectrometric method with area MDA-7100 diisocyanate monitors. During the three years the plant was in operation, short-term exposures did not exceed 5 ppb. The absence of elevated MDI levels during occasional spills was explained as either low volatility of MDI at room temperature, or by lack of monitors in close proximity to where the spills occurred.

Peak expiratory flow rate (PEFR) was performed in those workers (n=43) who reported at least one lower respiratory symptom of wheezing, cough, or shortness of breath, and in those workers with MDI-HSA specific antibodies. PEFR was performed in 23 workers free of symptoms and served as controls. Greater than 15 percent variability in PEFR between working and non-working conditions was found in three workers and considered a diagnosis of occupational asthma. One of the control workers also was diagnosed with occupational asthma, suggesting a false negative on the workers' questionnaire response and potential underestimation of the true prevalence of asthma.

Urticarial symptoms related to MDI sensitization were confirmed by elevated specific IgE levels and cutaneous reactivity to MDI-HSA in one worker. Lack of respiratory symptoms and known dermal exposure to MDI in this worker suggest the skin was the primary route of sensitization. Another worker had only elevated levels of serum specific IgE and IgG to MDI-HSA, but was free of respiratory symptoms.

The authors concluded that higher than normal exposures to MDI (i.e., >5 ppb) occurred during nonroutine activities in the workers with occupational asthma and MDI-related cutaneous anaphylaxis, and speculated that unpredictable exposure to MDI liquid and fumes occurring during maintenance or excessive heating of MDI-resin mixtures caused the observed reactions. In one of the workers, onset of asthmatic symptoms began 2 weeks after accidental exposure to a large MDI spill.

Sulotto et al. (1990)

End of work shift and end of work week effects on pulmonary function were examined in 27 asymptomatic polyurethane workers exposed to low levels of MDI. An equal number of clerks from the same factory with no exposure to MDI and without asthma were matched by age. The polyurethane workers were identified as having 14.0 years at work, but it is unclear if this time included isocyanate exposure for the entire duration. MDI sampling was carried out during the same time when lung function tests were performed. MDI concentrations ranged from 0.5 to 1 ppb and were analyzed using continuous tape monitoring. No significant differences in pulmonary function between the two groups were observed using a paired *t*-test over the course of a work day (Monday) or over the course of a work week (Monday before work to end of shift on Friday). Two-way ANOVA found reductions in FEV₁ and FEF₂₅₋₇₅ over the course of a work week, but the differences were related to smoking and not occupational exposure. The authors noted that 17 workers exhibiting isocyanate-induced asthma were removed from the facility sometime before the study began when TDI or mixtures of TDI-MDI were used. The authors concluded that short-term exposures to low levels of MDI do not result in respiratory changes.

Jang et al. (2000)

Jang et al. conducted a cross-sectional study of workers in Korean TDI and MDI manufacturing plants. A questionnaire was given and pulmonary function was performed on 20 workers exposed to MDI at a manufacturing plant, 44 workers exposed to TDI at a TDI manufacturing plant, and a control group consisting of 27 maintenance and field staff with no known exposure to the isocyanates. A total of 60 personal breathing zone samples were collected from the TDI and MDI workers during manufacturing processes using impingers. Sampling times were 30-60 min. The mean and maximum air concentrations of MDI was 1.3 and 6.4 µg/m³, and the mean and maximum air concentrations of TDI was 17.4 and 42.9 µg/m³. FEV₁ was comparable among all three groups. Airway hyperresponsiveness (AHR) was considered positive when a PC20 FEV₁ <16.0 mg/mL of methacholine was measured. By this criterion, AHR was greater (*p*<0.05) in MDI workers (20 percent) compared to TDI workers (4.7 percent). However, there was no difference when the MDI workers were compared to the control group. In both TDI and MDI workers that complained of respiratory symptoms, AHR was more prevalent (*p*<0.05). The authors observed no clear

evidence of sensitization, likely a result of the healthy worker effect (i.e., sensitized workers did not remain working at the plants).

Littorin et al. (2002)

Respiratory symptoms and biomarkers for isocyanate exposure were investigated in car industry workers (n=29) spraying or applying hot-melt glues containing PMDI onto flexible TDI polyurethane foam. Applying and heating of MDI-based glues were associated with biomarkers of inflammation in nasal lavage fluid (albumin, myeloperoxidase and neutrophils) and work-related symptoms of nasal irritation, including stuffiness, runny nose or sneezing. After work, workers who had complained of nasal irritation had higher levels of nasal inflammatory biomarkers than a control group of 15 office workers with no such history of exposure. However, biomarkers of TDI metabolites (mainly 2,6-toluene metabolites) in urine showed a stronger association with biomarkers of nasal inflammation than did metabolites of MDI. In addition, MDI metabolites were found in nasal lavage fluid of only two workers, whereas TDI metabolites were found in nasal lavage fluid of four workers. Finally, the presence of serum antibodies specific for TDI and MDI was associated with increased levels of nasal inflammatory biomarkers.

Littorin et al. (2002) had assumed the main exposure would be to fumes and gases emitted from the MDI-glue. However, a separate study conducted by the researchers showed that urinary TDI metabolites rose over a work shift when hot-melt glue was used, which was probably caused by thermodegradation of TDI-based polyurethane foam. The workers selected for this study were healthy workers (atopic workers were excluded) in order to reduce the “noise” of non-specific symptoms and signs.

Table 2 presents a summary of the findings from the occupational studies.

Table 2. Summary of occupational studies in workers exposed to MDI

Study	Study type, Industry & Exposure	Results
Pham et al., 1988	Longitudinal study Polyurethane production <20 ppb for all groups 118 directly exposed, 117 indirectly exposed and 83 unexposed workers	Initially, >asthma in directly exposed women (p<0.05) and lower predicted VC, FEV ₁ and DL _{CO} in directly exposed men with >60 mo of exposure After 5 yrs, no change in FEV ₁ , but observed increased loss of DL _{CO} in directly exposed workers (p<0.05)

Study	Study type, Industry & Exposure	Results
Petsonk et al., 2000	Prospective study Wood products plant Limited air measurements after the study below LOD. Possible dermal exposure	Of 178 workers, 12% had new onset of asthma after 2 years related to those working in high exposure areas ($p < 0.001$).
Johnson et al., 1985	Cross-sectional study foundry plant MDI Sampling: 86% < 0.05 ppb 9% > 5 & < 10 ppb 4% > 10 & < 15 ppb 1.5% > 15 ppb	Increased wet cough, lower FEV ₁ , FVC and FEF _{25-75%} compared to controls ($p < 0.05$). 13 of 66 workers had increased bronchial hyperreactivity by methacholine test. 6 of 9 tested by MDI provocation were positive for asthma. Cross contamination with silica dust.
Bernstein et al., 1993	Cross-sectional study in polyurethane workers 147 workers 96 controls ≤ 5 ppb during short-term measurements > 5 ppb during spills	3 workers and 1 control diagnosed with occupational asthma: $> 15\%$ variability in PEF, respiratory symptoms and/or MDI-HAS-specific antibodies
Sulotto et al., 1990	Cross-sectional study in polyurethane workers 27 asymptomatic workers 27 controls Exposures ranged from 0.5 to 1 ppb MDI	No end of work shift or end of work week changes in FEV ₁ or FEF ₂₅₋₇₅ in exposed workers. 17 workers with occupational asthma had been removed from plant prior to study
Jang et al., 2000	Cross-sectional study in MDI/TDI manufacturing plants. 20 MDI workers 44 TDI workers 27 control workers MDI exposure: mean 1.3 $\mu\text{g}/\text{m}^3$, maximum 6.4 $\mu\text{g}/\text{m}^3$	No difference in FEV ₁ among the groups. Airway hyperresponsiveness (PC ₂₀ FEV ₁ @ < 16.0 mg/mL of methacholine) greater in MDI and TDI workers ($p < 0.05$). Occupational asthma not seen, likely due to healthy worker effect.
Littorin et al., 2002	Cross-sectional study of workers applying/spraying glues with PMDI 29 non-atopic workers 15 control workers No air levels sampled Exposure estimated by MDI and TDI urinary metabolites	Increased biomarkers of inflammation in nasal lavage and increased work-related symptoms of nasal irritation in workers. Exposure better correlated with TDI exposure from polyurethane foam than with PMDI in glue

There are also case reports of neurological effects. In cases reported by Reidy and Bolter (1994), five individuals were occupationally exposed to MDI over a two-year span, and examined while exposure was ongoing (1 case), or up to 9 months after cessation of exposure (4 cases). The intensity and frequency of exposure was not reported and there was co-exposure to hydrocarbon solvent vapors. Subjective complaints included respiratory distress, headaches, forgetfulness, mood alterations, irritability, and difficulty concentrating. All subjects were diagnosed with isocyanate-induced occupational asthma and allergic rhinitis. Formal neuropsychological evaluations indicated that psychomotor, psychosensory, visuographic and language skills were largely intact. However, there was marked slowing in the rate of information processing, discrepancies in immediate recall of verbal versus nonverbal material, and deficiencies in learning ability. Complex, nonverbal abstract reasoning was impaired, and there was evidence of emotional distress in the form of depression, anxiety, and altered mentation. The authors concluded the data indicate compromised cognitive functions characteristic of CNS involvement, but do not clearly identify a single pattern of neuropsychological deficits associated with MDI exposure.

Several reports suggest that skin exposure to MDI in the workplace can increase the risk for sensitization and isocyanate-induced asthma (Bello et al., 2007). It was proposed that MDI skin exposure induces systemic sensitization, which then leads to occupational asthma following MDI inhalation exposure. Evidence largely results from settings in which known skin exposure occurs, but extremely low to non-detectable air levels of MDI are measured. In another occupational exposure study, MDI air concentrations were below the limit of detection in most breathing zone air samples, but detectable amounts of the MDI metabolite 4,4'-methylenedianiline (MDA) were found in nearly all urine samples (Kaaria et al., 2001). MDA is formed following acid hydrolysis of MDI metabolites in urine samples and is preferred for quantitative analysis. The presence of MDA in urine was explained, in part, by the long half-life of MDA in the body, and that exposure from previous days contributed to the urinary amount of metabolite. In addition, urinary MDA analysis was said to be a more sensitive and less arduous method than the established measurement of airborne MDI.

Case reports of MDI-exposed workers with allergic contact eczema on hands arms and face, and contact allergy (delayed dermal hypersensitivity) have been reported. Several cases of sensitization occurred after a few weeks or months of exposure at work (Estlander et al., 1992). In some cases the patient had MDI-induced asthma in addition to the contact eczema.

6.2 Chronic Toxicity to Infants and Children

No studies of the effects on infants and children of chronic exposure to MDI were located. However, early life exposure to MDI and other diisocyanates may occur

through inhalation and dermal contact with polyurethane products. Although detectable levels of free MDI emitted from bedding materials have not been found, Krone et al. (2003) applied semiquantitative tests (i.e., wipe test) for isocyanate to polyurethane products, including mattresses, mattress pads, sofa padding, carpet pads and pillows, and detected free isocyanate in these consumer products. When this bedding is used by infants, there is a potential for exposure of the developing immune system to MDI and other diisocyanates. In fact, among children with 12 or more wheezing attacks in the previous 12 months, independent associations have been reported between severe wheeze and the use of non-feather bedding, especially foam pillows (odds ratio 2.78; 95% C.I. 1.89 to 4.17) (Strachan and Carey, 1995). It has also been observed that there is a higher incidence of asthma among first children compared with their siblings (Karmaus and Botezan, 2002). This is consistent with the use of new polyurethane products (bedding, car seats, etc.) by the first child while siblings often use hand-me-downs whose isocyanate content has diminished over time.

These early life diisocyanate exposures may be significant since at birth, humans exhibit a dominant humoral, T_H2 , responsiveness (i.e., an atopic state). During the first few years of life, the T_H2 response converts to a more cellular (T_H1) immune response characteristic of the mature adult immune system. A delay in the transition from the predominant T_H2 pattern to the more balanced T_H1/T_H2 response allows an atopic T_H2 type response to persist longer, thus extending the period of vulnerability to environmental stressors and allergens, and increasing the likelihood of subsequent disease expression including asthma (Prescott et al., 1999).

In animal models of TDI-induced asthma, T_H2 cytokines were found to be decisive in the initial phase of asthma, in the priming of T_H2 cells, and in the permeation of eosinophils into the airway lumen (Matheson et al., 2005). In other mouse models of asthma, a T_H2 -dominated sensitization was best achieved via dermal exposure followed by tracheal exposure, while exposure exclusively by inhalation prompted more of a T_H1 type response (Ban et al., 2006). Thus if a T_H2 response predominates, whether by previous activation or due to an early developmental stage, subsequent respiratory exposure may result in a substantial pulmonary inflammatory response. Other examples of chemical stressors that can shift the immune response to a T_H2 pattern are diesel exhaust and ETS, both of which are strongly associated with asthma (Wang and Pinkerton, 2007).

6.3 Chronic Toxicity to Experimental Animals

In subchronic studies carried out by Reuzel et al. (1994b) prior to conducting a chronic exposure study, 10 rats/sex were exposed to PMDI concentrations of 0.35, 1.4 or 7.2 mg/m³ in the first study, and 20 rats/sex were exposed to 4.1, 8.4

or 12.3 mg/m³ PMDI in a second study. Exposure duration for both chamber studies was 6 hrs/day, 5 days/week for 13 weeks. The respiratory tract was the main target organ. Severe respiratory distress was observed in the 12.3 mg/m³ group. In treatment groups of 7.2 mg/m³ and above, accumulation of alveolar macrophages in the lungs was observed. The investigators also observed interstitial infiltration by macrophages at the highest two levels (8.4 and 12.3 mg/m³). At all concentrations in the second study, there was also significant accumulation of yellowish pigmented macrophages in mediastinal lymph nodes. In the nasal cavity, respiratory epithelial hyperplasia and olfactory epithelial atrophy were observed primarily in the two highest levels of the second study. For these pulmonary effects, the authors reported a NOAEL of 1.4 mg/m³.

The effects of chronic exposure to PMDI (approximately 50:50 monomeric:polymeric MDI) of 560 Wistar rats of both sexes was reported by Reuzel et al. (1994a). Animals were exposed to the PMDI mixture for 6 hr/day, 5 days/week for one or two years. The mean exposure concentrations were 0, 0.19, 0.98 and 6.03 mg/m³, with mass median aerodynamic diameters of 0.68, 0.70, and 0.74 µm, respectively. After both one and two years of exposure, there was significant ($p < 0.01$) accumulation of macrophages with yellow pigment in the lungs and the mediastinal lymph nodes at the highest dose (6.03 mg/m³), and at 0.98 mg/m³ after two years of exposure.

Alveolar duct epithelialization as well as fibrosis of tissues surrounding the macrophage accumulations occurred in the 0.98 and 6.03 mg/m³ exposure groups (Table 3). Localized alveolar bronchiolization was seen in the 6.03 mg/m³ group. The time sequence of the spectrum of pulmonary changes indicates that recurrent alveolar wall damage by PMDI and/or polymeric-containing alveolar macrophages leads to alveolar epithelialization.

In the nasal cavity, minimal to moderate olfactory epithelial rearrangement was observed in all treatment groups, but was significant only at the highest dose. Basal cell hyperplasia and Bowman's gland hyperplasia were significant ($p < 0.05$) in males at the two highest dose levels after two years (Table 3). In females, basal cell hyperplasia was significant ($p < 0.01$) only at the highest dose, and Bowman's gland hyperplasia was not statistically significantly increased at any dose. The authors concluded that these data indicate a LOAEL of 0.98 mg/m³, and a NOAEL of 0.19 mg/m³ for the noncancer respiratory tract effects.

Table 3. Incidences of primary microscopic findings in the lungs and nasal cavity of male and female Wistar rats exposed to PMDI for 2 years (Reuzel et al., 1994)^a

	Males				Females			
	0	0.19	0.98	6.03	0	0.19	0.98	6.03
Exposure concentration (mg/m ³)	0	0.19	0.98	6.03	0	0.19	0.98	6.03
Number of lungs examined	60	60	60	60	59	60	60	59
Lungs								
Localized fibrosis	1	0	9*	44**	0	0	4	48**
Minimal	0	0	7	5	0	0	3	8
Mild	1	0	2	22	0	0	1	23
Moderate	0	0	0	17	0	0	0	14
Severe	0	0	0	0	0	0	0	3
Alveolar duct epithelialization	1	0	8*	54**	0	0	8*	57**
Localized alveolar bronchiolization	1	1	2	12**	2	3	3	14**
Nasal Cavity								
Basal cell hyperplasia	14	13	26*	32**	4	8	8	49**
Olfactory epithelial degeneration	6	9	15	25*	8	16	10	20*
Bowman's gland hyperplasia	0	2	9**	17**	2	6	8	8

^a Values marked with asterisks differ significantly from control (*p<0.05, **p<0.01)

In a separate chronic study by Hoymann et al. (1998), female Wistar rats were subjected to whole-body exposures to monomeric MDI aerosols (0.23, 0.70, and 2.05 mg/m³; MMAD 1.03-1.06 µm) for 17 hours/day, 5 days/week for up to 24 months. Chronic exposure to MDI caused a time and dose-dependent deterioration of lung function. Lung function was assessed after 6, 12, 17 and 20 months, with histological evaluations after 12 and 24 months. At all time points, the highest exposure (2.05 mg/m³) caused a significant decrease in maximum mid-expiratory flow (MMEF), and forced expiratory flows (FEF) at 10, 25, and 50%, but not 75% of forced vital capacity. This indicates a significant increase in flow resistance in the small peripheral airways, but not the large airways. With the longer 12 and 17 month exposures, the decrements in these flow measures were seen at the lower doses as well. After 12 and 17 months at the high dose, the CO diffusion test showed a reduction in diffusion through the alveolar-capillary membrane.

Lung weights in the the high dose group were increased after 3, 12, and 20 months in the Hoymann et al. study, and correlated well with the histopathological findings (see Table 4) of alveolar and bronchiolar hyperplasia presented in a separate report by Ernst et al. (1998) of the same study. In bronchoalveolar lavage fluid (BALF) obtained at these same time points, elevated hydroxyproline indicated increased collagen metabolism that correlated well with histopathological findings of interstitial fibrotic lesions in the lungs.

Examination of BALF also showed an inflammatory reaction, with increased numbers of lymphocytes, at all time points at the highest dose.

In the nasal cavity, MDI-related lesions included degeneration and focal squamous metaplasia of the olfactory epithelium (Ernst et al., 1998). MDI-related focal squamous metaplasia was observed in the larynx, and lymphoid hyperplasia and accumulation of particle laden macrophages were observed in the lung associated lymph nodes. These lesions were not quantified in the histopathology results.

The results reported by Hoymann et al. and Ernst et al. are consistent with histopathologically determined dose-dependent interstitial and peribronchiolar fibrosis causing fibrotic thickening of walls of peripheral bronchioles and narrowing of small airways. The decline in lung function started before 6 months of exposure, increased through 12 months, and increased more slowly through 17 months. Measures of MMEF and FEF at 12 months suggest a LOAEL of 0.2 mg/m³, the lowest dose used.

Table 4. Incidences of MDI-related pulmonary lesions in female rats exposed to monomeric MDI for 2 years (Ernst et al. 1998)^a

	MDI Concentration (mg/m ³)			
	0	0.2	0.7	2.1
Number of lungs examined	80	80	80	80
Peribronchiolar and interstitial fibrosis	4	51***	73***	77***
Very slight	1	36***	29***	7
Slight	3	15**	43***	45***
Moderate	0	0	1	25***
Bronchiolo-alveolar hyperplasia	3	6	14*	41***
Very slight	2	2	7	10*
Slight	1	4	7	24***
Moderate	0	0	0	5
Severe	0	0	0	2
Alveolar cell hyperplasia	2	8	12*	21***
Very slight	0	1	0	0
Slight	2	4	5	6
Moderate	0	2	5	8**
Severe	0	1	2	7*

^a Values marked with asterisks differ significantly from control (*p<0.05, **p<0.01, ***p<0.001)

Table 5 presents the benchmark concentration (BMC) modeling results of the respiratory system endpoints from the histopathology results presented in Tables 3 and 4. OEHHA performed BMC modeling using U.S EPA benchmark dose

software, version 2.3.1 (U. S. EPA, 2012). The BMC_{05} represents the 5% response rate for the endpoint and the $BMCL_{05}$ represents the 95% lower confidence limit of the dose producing a 5% response rate (the BMC_{05}). Using the OEHHA BMC modeling approach outlined in the OEHHA Noncancer TSD (OEHHA, 2008), the $BMCL_{05}$ is used as the point of departure for noncancer risk assessment. BMC_{05} and $BMCL_{05}$ were derived from the model that provided the best visual and statistical fit to the data among the group of models, particularly in the low dose region where the BMC_{05} resides. Following U.S. EPA guidelines, we chose the model with the lowest AIC (Akaike information criterion) in instances where various acceptable model fits to the data were similar.

Table 5. BMCs and BMCLs for the main respiratory system lesions in the 2-year MDI inhalation exposure study in rats (Reuzel et al., 1994).

Respiratory System Endpoint	Model	BMC_{05} (mg/m^3)	$BMCL_{05}$ (mg/m^3)
Reuzel et al. (1994)			
Localized fibrosis	Log-probit	0.721	0.554
Alveolar duct epithelialization	Log-probit	0.705	0.569
Nasal basal cell hyperplasia	Weibull	0.313	0.253
Nasal olfactory epithelial degeneration	Log-logistic	0.854	0.549
Ernst et al. (1998)			
Peribronchiolar & interstitial fibrosis	ND*	ND	ND
Bronchiolo-alveolar hyperplasia	Probit	0.419	0.351
Alveolar cell hyperplasia	Log-logistic	0.324	0.213

* Not determined. An adequate fit to the data could not be found with the available BMD models.

The incidence data for peribronchiolar and interstitial fibrosis from the Ernst et al. study in Table 5 could not be adequately modeled with any of the available BMD models, although this was the most sensitive endpoint of MDI exposure. This was likely due to the steep dose-response between the control and low dose groups.

In consultation with the pathologists that examined the lungs in the Reuzel et al. and Hoymann et al./Ernst et al. chronic studies (from here on simply referred to as the Hoymann et al. study), Feron et al. (2001) re-examined the histopathological data of the lung sections in female rats from both studies. The nomenclature and grading schemes used to describe histopathological changes in lung tissue were harmonized in a joint effort between the pathologists involved in the original studies and an independent reviewing pathologist not involved in the original reading of the slides. In general, many similarities were found in the toxicological profiles from the review of the two studies. The differences found were ascribed as probably a consequence of the experimental protocols used rather than due to differences in intrinsic toxicity of the test materials. Specifically, Reuzel et al. employed a 6 hr/day exposure protocol, while Hoymann et al. employed an 18 hr/day (corrected upwards from 17 hrs/day as presented in the

original study) exposure protocol. Also, Reuzel et al. exposed rats to PMDI and Hoymann et al. employed monomeric MDI.

Major dose-related microscopic lung lesions quantified included interstitial fibrosis and bronchiolo-alveolar hyperplasia (Table 6). Subclassifications of bronchiolo-alveolar hyperplasia included alveolar- and bronchiolar-type hyperplasia, and mixed- and flat-type hyperplasia. Appearance of mixed- and flat-type hyperplasia was also recorded, but was irregular and did not show dose-related trends (data not shown).

Table 6. Reexamination by Feron et al. (2001) of the incidences of microscopic findings in the lungs of female rats exposed to PMDI (Reuzel et al., 1994) and MDI (Hoymann et al., 1998)*.

	Reuzel et al. Study PMDI				Hoymann et al. Study MDI			
	0	0.19	0.98	6.03	0	0.23	0.7	2.05
Exposure concentration (mg/m ³)	0	0.19	0.98	6.03	0	0.23	0.7	2.05
Number of lungs examined	59	60	60	59	80	80	80	80
Bronchiolo-alveolar hyperplasia	11	10	25	59	8	16	27	53
Alveolar-type hyperplasia	7	5	8	30	2	11	13	29
Grade 1	3	0	2	5	1	8	4	12
Grade 2	1	4	2	8	1	2	5	3
Grade 3	2	1	3	9	0	1	2	7
Grade 4	1	0	1	6	0	0	1	6
Grade 5	0	0	0	2	0	0	1	1
Bronchiolar-type hyperplasia	0	1	12	59	2	8	20	42
Grade 1	0	0	9	17	2	6	18	29
Grade 2	0	1	3	37	0	2	2	10
Grade 3	0	0	0	5	0	0	0	3
Interstitial fibrosis	2	2	19	59	10	63	77	79
Grade 1	2	2	19	7	7	36	26	2
Grade 2	0	0	0	44	3	25	41	29
Grade 3	0	0	0	7	0	2	9	34
Grade 4	0	0	0	1	0	0	1	14

* Statistical analysis of lesion incidences were not conducted

Calculated as cumulative dose, the high exposure group from each study was similar (17,728 mg-hr/m³ for the Reuzel. et al study and 17,575 mg-hr/m³ for the Hoymann study). However, Table 6 shows that there was a higher incidence and greater extent of proliferative epithelial changes in rats from the Reuzel study at this dose. Feron et al. (2001) attributed this difference, in part, to the better survival of the rats in the Reuzel study. Average survival in the Reuzel rats was 700 days, whereas average survival in the Hoymann rats was 518 days. Feron et al. also surmised that there was a higher local dose and tissue deposition

during exposure in the high dose rats from the Reuzel study, which received 6.03 mg/m³ PMDI versus 2.05 MDI in the high dose Hoymann rats.

The incidence and/or degree of interstitial fibrosis were clearly higher in the Hoymann study than in the Reuzel study (Table 4). Feron et al., (2001) postulated that the stronger fibrotic response was probably related to the longer daily exposure period for the rats in the Hoymann study, which were exposed 18 hr/day versus 6 hr/day in the Reuzel study. The Hoymann rats experienced much longer deposition of freshly inhaled MDI in the alveolar duct region, most likely leading to greater damage of the local tissue, with a subsequent shorter (overnight) recovery period. The more pronounced interstitial fibrosis in the Hoymann study cannot be related to longer survival or higher exposure concentration since, as already noted, survival was longer and exposure concentration was higher in the high-dose group of the Reuzel study (Feron et al., 2001).

Based on the histopathological findings of the 2-year exposure studies, Feron et al. (2001) estimated a NOAEL for MDI exposure essentially using a NOAEL/LOAEL approach. Due to the mild tissue effects at the low dose in the Hoymann et al. and Ernst et al. studies and a no adverse effect at the low dose in the Reuzel study, Feron et al. suggested that 6 hr exposures to 0.23 mg/m³ as a NOAEL for both MDI or PMDI.

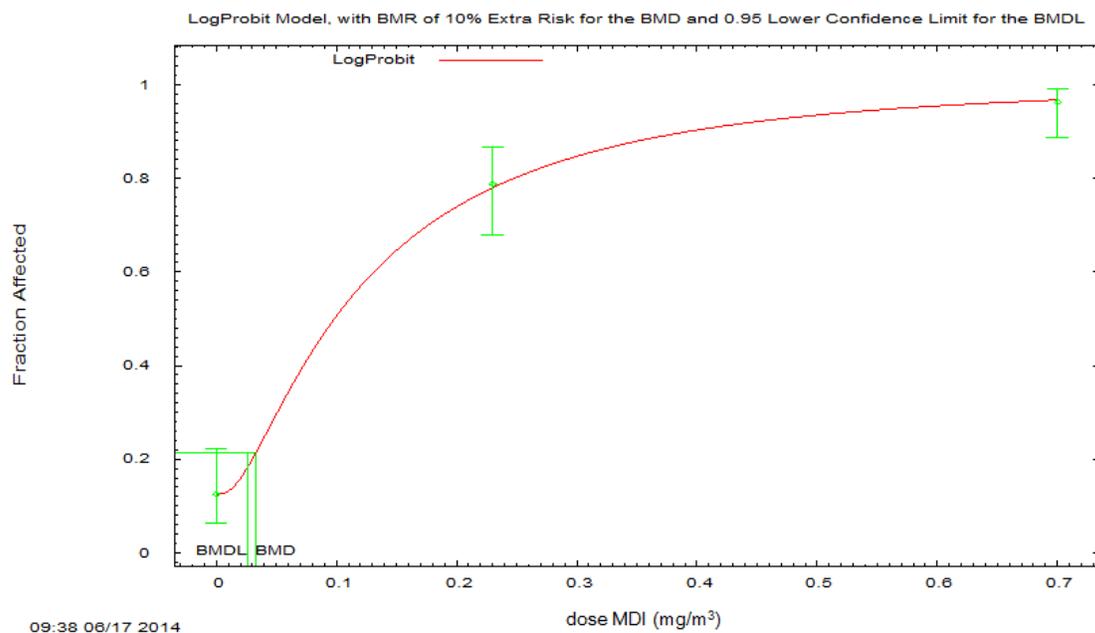
From the total incidence data presented in Table 6, dichotomous benchmark dose modeling was performed by OEHHA on the reexamination of respiratory endpoints by Feron et al. (2001). Table 7 shows the best model fit of the data and the resulting BMCs and BMCLs. Interstitial fibrosis was the most sensitive indicator of MDI exposure with the 18 hrs/day, 5 days/week exposure protocol used in the Hoymann study, and bronchiolo-alveolar hyperplasia was the most sensitive indicator of PMDI exposure with the 6 hrs/day, 5 days/week exposure protocol used in the Reuzel study. An appropriate model fit for the data was achieved with a BMC₁₀ and dropping the high dose group (Figure 4). The steep dose-response from control to low dose for the endpoint could not be adequately modeled with a BMC₀₅, likely due to the 5% response level being beyond the limit of sensitivity. US EPA (2012) generally recommends using a BMC10 for analysis unless enough data is near the observable range for the 5% response rate. In addition, a better model fit could be obtained if the high dose group was removed. The scaled residual exceeded 2 for nearly all models when the data was modeled with the high dose included. This led to a poor p-value (p<0.1) and a failure for the goodness-of-fit test. Dropping the high dose group can be done in BMD modeling if the two highest dose groups are at or near 100% response, as was the case for incidence of interstitial fibrosis, resulting in the high dose group providing little useful information for estimation of the BMCL₀₅ at the low end of the dose-response curve.

Table 7. BMCs and BMCL₀₅s for the major respiratory system endpoints in the 2-year MDI inhalation exposure studies in rats (reanalysis of Reuzel and Hoymann findings by Feron et al., 2001).

Respiratory System Endpoint	Model	BMC (mg/m ³)	BMCL ₀₅ (mg/m ³)
Interstitial Fibrosis (Hoymann)	Log-probit	0.0326*	0.0256
Bronchiolar Hyperplasia (Hoymann)	Weibull	0.144	0.116
Alveolar Hyperplasia (Hoymann)	Log-Logistic	0.203	0.143
Bronchiolo-alveolar hyperplasia (Hoymann)	Quantal linear	0.109	0.087
Interstitial Fibrosis (Reuzel)	Logistic	0.383	0.294
Bronchiolar Hyperplasia (Reuzel)	Multistage	0.416	0.233
Alveolar Hyperplasia (Reuzel)	Logistic	1.154	0.931
Bronchiolo-alveolar hyperplasia (Reuzel)	Multistage	0.376	0.118

* Best fit for the interstitial fibrosis data was obtained by modeling with a BMC₁₀ and dropping the high exposure group. All other endpoints in this table were modeled with a BMC₀₅ and with all four exposure groups.

Figure 4. Log-probit model (BMC₁₀) fit to the 2-year MDI Hoymann study for interstitial fibrosis in rats.



BMC modeling of the microscopic findings was also run using continuous models supplied by U.S. EPA in their BMD software (U. S. EPA, 2012). A weighted average approach was used to convert the severity and incidence data in Table 6 to a mean and standard deviation for each dose group. This was achieved by assigning each severity category a number: Grade 0 = zero, Grade 1 = one, Grade 2 = two, Grade 3 = 3, and Grade 4 = 4. Only limited success was attained in fitting continuous models to the data (Table 8).

Table 8. BMCL_{1 SD} and BMCL_{0.5 SD} continuous modeling results for the major respiratory system endpoints in the 2-year MDI inhalation exposure studies in rats (reanalysis of Reuzel and Hoymann findings by Feron et al., 2001).

Respiratory Endpoint	Dichotomous BMCL ₀₅	Continuous BMCL _{1 SD}	Continuous BMCL _{0.5 SD}
Hoymann et al.			
Pulmonary Fibrosis	0.0256 (log-probit)	ND	ND
Bronchiolar Hyperplasia	0.116 (Weibull)	0.270 (Exponential M4)	0.130 (Exponential M4)
Alveolar Hyperplasia	0.143 (log-logistic)	0.329 (Polynomial)	0.158 (Polynomial)
Reuzel et al.			
Pulmonary Fibrosis	0.294 (logistic)	1.026 (rho=0) (Exponential M4)	0.524 (rho=0) (Exponential M4)
Bronchiolar Hyperplasia	0.223 (multistage)	ND	ND
Alveolar Hyperplasia	0.931 (logistic)	5.242 (rho=0) (Exponential M2/3)	3.613 (rho=0) (Exponential M2/3)

ND: Not determined. An adequate fit to the data could not be found with the models in the BMD program, or not enough information was provided for an analysis with continuous models.

Models that gave acceptable values for fit and the lowest AIC are shown in parenthesis in Table 8 for each respiratory system endpoint. We specified a risk of both 1 and 0.5 estimated standard deviation from the control mean as benchmarks (i.e., the BMC). The 95th lower confidence limit on the BMC in Table 8 represents the potential point of departure for a REL derivation (i.e., the BMCL_{1 SD} and BMCL_{0.5 SD}). Included in Table 8 are the BMCL₀₅ based on dichotomous modeling for comparison to the continuous modeling results. Note that the results of the continuous model using 0.5 SD are generally closer to those of the dichotomous models than the continuous modeling results using 1.0 SD.

Acceptable continuous model fits to the data could not be estimated for pulmonary fibrosis from the Hoymann et al. study and for bronchiolar hyperplasia from the Reuzel et al. study. In other cases, the continuous BMD modeling output recommended non-homogenous modeling (rho≠0) be used. However, non-homogenous modeling did not provide an adequate fit to the data, so the homogenous modeling results are presented in Table 8 for two respiratory system endpoints (i.e., pulmonary fibrosis and alveolar hyperplasia from the Reuzel et al. study). Incidence data was only supplied for bronchiolo-alveolar hyperplasia, so continuous modeling could not be performed on this endpoint.

Feron et al. (2001) also compared the physical properties of MDI used in both chronic exposure studies. Chemical analysis by high pressure liquid chromatography (HPLC) of the test atmospheres carried out in both the Reuzel and Hoymann studies revealed that, at the lowest exposure concentration, a certain percentage of monomeric MDI was present as vapor. The estimated

saturation concentrations were 100 and 50 $\mu\text{g}/\text{m}^3$ at 25°C for monomeric and polymeric MDI, respectively. The inhaled dose of MDI aerosol was calculated by subtracting the saturation vapor concentration from the total concentration. At the lowest dose (0.19 mg/m^3) in the Reuzel et al. study, 72 percent of the total inhaled dose of PMDI is expected to be in the aerosol phase. At the lowest dose (0.23 mg/m^3) in the Hoymann et al. study, 56 percent of the total inhaled dose of MDI is expected to be in the aerosol phase. For regional deposition of the aerosol forms in the upper respiratory tract, Feron et al. calculated about 17 and 20 percent nasal deposition of the inhaled aerosol for MDI in the Hoymann study and PMDI in the Reuzel study, respectively.

6.4 Toxicogenomics

Even though diisocyanates are one of the most common causes of occupational asthma, only 5-15 percent of exposed workers develop the disease. Thus, genetic predisposition has been implicated in the susceptibility to occupational asthma by MDI and other diisocyanates. A number of gene variants have been reported to be associated with increased sensitivity to the disease in workers (Table 9), which suggests that diisocyanate-induced asthma represents a complex disease phenotype determined by multiple genes. Examples of genes include, but are not limited to, genes involved in immune regulation, inflammatory regulation, and antioxidant defense (Choi et al., 2009; Yucesoy and Johnson, 2011; Yucesoy et al., 2012). The anticipated goal of genetic association studies is to provide more accurate information on interindividual variability, thereby contributing to better protect sensitive human populations and to establish more accurate exposure limits in the workplace. No studies examined only MDI-exposed workers. The summarized toxicogenomic studies below include mixed cohorts exposed to TDI, MDI and/or HDI, or to TDI only.

A case-control study was conducted by Yucesoy et al., (2012) to investigate whether genetic variants of antioxidant defense genes are associated with increased susceptibility to diisocyanate-induced asthma (DA). The study population consisted of 353 Caucasian French-Canadians in three groups: 95 workers with specific inhalation challenge confirmed DA; 116 symptomatic diisocyanate workers with a negative specific inhalation challenge; and 142 asymptomatic exposed workers. Diisocyanate exposures were to HDI, MDI and/or TDI. Antioxidant defense gene variations for superoxide dismutase, glutathione-S-transferase and epoxide hydrolase and their interactions were found to contribute to DA susceptibility (Table 9).

In a study population of 109 cases of workers with DA (HDI-, MDI- and TDI-exposed) and 73 exposed non-symptomatic controls, polymorphisms in glutathione-S-transferase genes were examined to look for associations with DA (Piiirila et al., 2001). Lack of the GSTM1 gene (null) was found to be associated with increased risk of DA. In a later study on the same worker group, Wikman et al., (2002) investigated the possible role of *N*-acetyltransferase (NAT) genotypes in the development of DA. Positive associations for increased DA were found

with slow acetylator genotypes and genotype combinations with a glutathione-S-transferase (GSTM1 null) genotype.

The human leucocyte antigen (HLA) class II molecules are also thought to be involved in the development of the immune response to diisocyanates. HLA class II molecules are encoded by genes located within the major histocompatibility complex and present antigens from outside of the cell to T-lymphocytes. These particular antigens stimulate the multiplication of T-helper cells, which in turn stimulate antibody-producing B-cells to produce antibodies to that specific antigen. Mapp et al. (2000) examined the distribution of markers (DQA, DQB and DRB) for HLA class II genes in European Caucasians (67 TDI-exposed workers with DA, 27 asymptomatic TDI worker controls, and 101 normals). The frequencies of DQA1*0104 and DQB1*0503 were significantly increased in asthmatic subjects. DQB1*0503 was also more frequent among asthmatic subjects compared with normal subjects.

Kim et al. (2006) evaluated a Korean population for associations of HLA class I and II alleles with TDI-induced asthma. Statistical significance was not found with single allele analysis. However, two and three locus haplotype analysis showed several significant alleles as potential susceptible markers. The authors identified the HLA haplotype DRB1*15-DPB1*05 as the most useful marker for predicting development of TDI-induced DA in the Korean population.

CTNNA3 (alpha-T catenin) is a key protein of the adherence junctional complex in epithelial cells and plays an important role in cellular adherence. The function of CTNNA3 in diisocyanate-induced asthma is not known, but it has been shown that decreased expression of CTNNA3 may lead to increased susceptibility to diisocyanate effects and contribute to development of DA (Bernstein et al., 2013). A mainly French-Canadian Caucasian study population including 132 workers (TDI-, MDI- or HDI-exposed) with DA, 131 symptomatic workers with a negative challenge for DA, and 147 asymptomatic workers were examined to determine if genetic variants of CTNNA3 genes are associated with increased susceptibility to DA. CTNNA3 minor allele homozygotes of rs7088181 and rs10762058 SNPs were at increased risk for DA compared to the asymptomatic control workers, but not symptomatic workers with a negative challenge. These same CTNNA3 single-nucleotide polymorphisms (SNPs) were also significantly associated with TDI-induced asthma in a group of 84 Korean workers with DA compared to 263 normal controls (Kim et al., 2009).

Sixty-two workers with DA and 75 diisocyanate workers negative for DA were analyzed for SNPs associated with the immune response genes IL4RA, IL-13, and CD14 (Bernstein et al., 2006). The T_H2 cytokines IL-4 and IL-13 play key roles in B-cell IgE isotype class switching and are believed to at least partially determine expression of airway inflammation and allergic disease. No associations were found with individual SNPs and DA when all diisocyanate workers (TDI-, MDI- and HDI-exposed) were considered. When only HDI-

exposed workers were considered (34 with DA, 62 negative for DA), associations with immune response genes and DA were found. The strongest associations were for the two-genotype variation combination IL4RA (150V) II and CD14 (C159T) CT, and the three-genotype variation combination IL4RA (150V) II, IL13 (R110Q) RR, and CD14 (C159T) CT.

Table 9. Variability in Observed Odds Ratio (OR) or *p* Value for Significant Genotype Variation Associations and Increased Susceptibility for Diisocyanate-Induced Asthma

Reference	Ratio or <i>p</i> value	Genetic associations for DA
Yucesoy et al., 2012	OR=2.70 ^a (95%CI 1.38-5.27) <i>p</i> =0.004	SOD2 (rs4880) superoxide dismutase single-nucleotide polymorphism (SNP) Ala→Val substitution on SOD2 gene that decreases the activity of SOD2
	OR=6.10 ^a (95%CI 1.31-28.4) <i>p</i> =0.021	GSTP1 (rs762803) glutathione-S-transferase SNP of unknown functional consequence
	OR=7.34 ^a (95%CI 2.04-26.5) <i>p</i> =0.002	GSTM1*EPHX1 (rs2854450) copresence of glutathione-S-transferase (GSTM1) deletion and minor allele for epoxide hydrolase (EPHX1 rs2854450)
	OR=8.55 ^a (95%CI 1.05-69.9) <i>p</i> =0.045	EPHX1 (rs2740168)*EPHX1 (rs1051741) copresence of two EPHXs, rs2740168 variant and a variation (rs1051741) that reduces enzyme activity
	OR=10.36 ^b (95%CI 1.47-72.96) <i>p</i> =0.019	EPHX1 (rs1051741) epoxide hydrolase minor allele
	OR=6.22 ^b (95%CI 1.95-19.82) <i>p</i> =0.002	EPHX1 (rs2740171) epoxide hydrolase SNP minor allele
Piirila et al., 2001	OR=1.89 (95%CI 1.00-3.52)	GSTM1 (null) gene lacks enzyme activity (59 cases and 29 controls with TDI, MDI or HDI exposure)
Wikman et al., 2002	OR=7.77 (95%CI 1.18-51.6)	NAT1 gene polymorphism for slow acetylation. TDI-exposed only (23 cases, 8 controls)
	OR=4.53 (95%CI 1.76-11.6) <i>p</i> =0.040	GSTM1 (null)*NAT1 copresence (43 cases and 20 controls with TDI, MDI or HDI exposure)
Mapp et al., 2000	<i>P</i> =0.005	HLA DQA1*0104 - 16 of 23 cases (23.9%), 0 of 10 asymptomatics (0%)
	<i>P</i> =0.009	HLA DQB1*0503 - 14 of 23 cases (20.9%), 0 of 10 asymptomatics (0%)
	<i>P</i> =0.027	HLA DQB1*0503 - 14 of 23 cases (20.9%), 9 of 30 normals (8.9%)

Reference	Ratio or <i>p</i> value	Genetic associations for DA
Kim et al., 2006	<i>P</i> =0.001 (cases vs. asymptomatics) <i>P</i> =0.003 (cases vs. normals)	HLA DRB1*15-DPB1*05 - 10.6% in cases (n=110), 0% in asymptomatic controls (n=94), and 2.5% in normals (n=190).
Bernstein et al., 2013	OR=9.05 (95%CI 1.69-48.54) <i>p</i> =0.01	CTNNA3 (rs7088181) – SNP minor allele homozygote (130 cases, 147 asymptomatic control workers)
	OR=6.82 (95%CI 1.82-14.88) <i>p</i> =0.002	CTNNA3 (rs10762058) – SNP minor allele homozygote (130 cases, 147 asymptomatic control workers)
Bernstein et al., 2006	OR=5.2 (95%CI 1.65-28.24) <i>p</i> =0.008	IL4RA (150V) II and CD14 (C159T) CT HDI workers with DA 39% vs 11% among DA-negative workers
	OR=6.4 (95%CI 1.57-26.12) <i>p</i> =0.01	IL4RA (150V) II, IL13 (R110Q) RR, and CD14 (C159T) CT HDI workers with DA 24% vs 5% among DA-negative workers

7. Developmental and Reproductive Toxicity

To examine the prenatal toxic effects of monomeric MDI aerosols, Buschmann et al. (1996) exposed pregnant Wistar rats to 0, 1, 3, and 9 mg/m³ for 6 hours per day on gestational days 6 to 15. At sacrifice on gestational day 20, lung weights were significantly increased in the high dose group (*p* < 0.01), as were the number of litters with fetuses displaying asymmetric sternebra (*p* < 0.05) (Table 10). Treatment reportedly had no effect on maternal weight gain, number of corpora lutea, implantation sites, pre- and post-implantation loss, fetal and placental weight, gross and visceral anomalies, and degree of ossification. In the mid-dose range, slight deviations were observed in numbers of fetuses with dilated ureters, accessory lumbar ribs and incomplete ossification of sacral vertebral centers. Maternal food consumption decreased at 1 and 3 mg/m³ at some time points, but did not affect weight gain. Maternal lung weights were increased at 9 mg/m³. A slight but significant increase in litters with fetuses displaying asymmetric sternebra(e) was observed in the 9 mg/m³ group (Table 10).

Table 10. Litters with Asymmetric Sternebra (Buschmann et al., 1996)

MDI (mg/m ³)	Litters (total)	Asymmetric Sternebra			% Litters with skeletal anomalous fetuses
		No. of litters	% of litters	% of fetuses	
0	25	2	8	2.3	60
1	26	7	27	4.6	61.5
3	25	5	20	2.8	64
9	23	10*	43	5.5	69.6

* *p* < 0.05

The authors noted asymmetric sternebrae is a minor variation and is common in the strain of rat used and in rats generally. The authors also reported that the observed incidence of this variation was within the normal historical range. The percent of fetuses with skeletal anomalies and percent of litters with skeletal anomalous fetuses were unaffected by MDI. Buschmann et al. concluded that a substance-induced effect on the sternebra cannot be ruled out at 9 mg/m³ and suggest 3 mg/m³ as a NOAEL for embryotoxicity.

The prenatal effects of aerosols of the polymeric form of MDI were also examined in Wistar rats exposed on gestational days 6 through 15 to 0, 1, 4, and 12 mg /m³ for 6 hours per day (Gamer et al., 2000). Maternal toxicity was clearly evident at the highest dose with significantly reduced body weight gain during pregnancy ($p < 0.01$) and, at sacrifice on day 20, significantly reduced organ and carcass weights. Fetal body weight per litter and placental weights per litter were also reduced at this dose ($p \leq 0.01$ and $p \leq 0.05$, respectively). Significant fetal toxicity manifested primarily at the highest dose and as skeletal malformations ($p < 0.01$). These included irregularly shaped sternebrae, bipartite sternebrae, and incomplete ossified vertebral bodies. The number of affected fetuses per litter was increased at the highest concentration of 12 mg/m³ for skeletal variations and skeletal malformations. All of these findings were above the historical control range for this strain of rat at 12 mg/m³ only. Measured in terms of affected fetuses per litter, the incidences of total fetal variations were significantly increased in all exposed groups (Table 11). However, only at the high concentration was the incidence above the historical control range. The authors attribute the total incidence of fetal variations to an unexpectedly low incidence in the control group, and suggest a NOAEL of 4 mg/m³ for maternal and developmental toxicity of PMDI.

Table 11. Incidence of total fetal variations (Gamer et al., 2000)

Total Variations	PMDI (mg/m ³)				Historical Control Range (%)
	0.0	1.0	4.0	12.0	
Fetuses: no. (%)	84 (25.0)	118 (35.0)	116 (33.6)	109 (39.1)	21.0-44.0
Litters: no. (%)	23 (92.0)	24 (100.0)	23 (95.8)	21 (100.0)	88.0-100.0
Affected fetuses/litter	24.6±15.7	34.7±15.6*	33.4±16.7*	40.0±12.1**	20.7-35.2

* $p < 0.05$; ** $p < 0.01$

8. Derivation of Reference Exposure Levels

8.1 MDI Acute Reference Exposure Level

<i>Study</i>	Pauluhn, 2002
<i>Study population</i>	Female Wistar rats
<i>Exposure method</i>	Whole body inhalation to PMDI 0, 0.7, 2.3, 8 and 20 mg/m ³
<i>Continuity</i>	Single exposure
<i>Duration</i>	6 hr
<i>Critical effects</i>	Increased total protein in BALF three hours post-exposure
<i>LOAEL</i>	0.7 mg/m ³ (0.068 ppm)
<i>NOAEL</i>	Not observed
<i>BMCL₀₅</i>	No acceptable model fit
<i>Time-adjusted exposure</i>	4.2 mg/m ³
<i>Human equivalent concentration</i>	9.45 mg/m ³ RGDR =2.25 for pulmonary region effects
<i>LOAEL uncertainty factor</i>	6
<i>Interspecies uncertainty factor</i>	
<i>Toxicokinetic (UF_{a-k})</i>	2
<i>Toxicodynamic (UF_{a-d})</i>	√10
<i>Intraspecies uncertainty factor</i>	
<i>Toxicokinetic (UF_{h-k})</i>	√10
<i>Toxicodynamic (UF_{h-d})</i>	10
<i>Cumulative uncertainty factor</i>	1200
<i>Reference Exposure Level</i>	8 µg/m ³ (0.8 ppb)

Reference Exposure Levels (RELs) are based on the most sensitive and relevant health effects reported in the medical and toxicological literature. Acute RELs are levels at which infrequent one-hour exposures are not expected to result in adverse health effects (OEHHA, 2008). The acute REL is based on increased total protein in BALF, which is an indicator of pulmonary epithelial injury and/or compromised function of pulmonary epithelium. Six-hour exposure to PMDI in rats resulted in increased total protein at the lowest exposure concentration (0.7 mg/m³) three hours post-exposure. BMC continuous modeling resulted in an acceptable fit for increased total protein at one day post-exposure, but not at three hours post-exposure when the effect was most pronounced.

Although the acute REL is based on cellular epithelial effects of PMDI in the pulmonary region of the lung, sensitization resulting in asthma is likely a greater concern for human exposure to MDI. However, Pauluhn and Poole (2011) observed in rats that allergic responses resulting from inhalation of PMDI appear to be linked with pulmonary irritation of the lower airways. Their *C × t* study showed that the dose that triggers an elicitation response in the rat asthma model is slightly below that causing acute pulmonary irritation in naïve rats. This

finding suggests that avoidance of MDI exposures that result in cellular pulmonary effects may avoid asthmatic responses in non-sensitized asthmatic individuals. By extension, acute exposures to PMDI below the threshold for pulmonary irritation may also avoid acute respiratory sensitization to MDI.

A time extrapolation from the 6 hr exposure to 1 hr was used applying Haber's Law $C^n \times T = K$ with an "n" = 1 based on the $C \times t$ study by Pauluhn (2002). This study showed that the magnitude of BALF protein matched the exposure intensity over the range of 3.4 to 58.1 mg PMDI/m³ and exposure durations of 6 hr to 23 min, indicating equal dependence on changes in concentration and duration of exposure. OEHHA notes that an assumption is made in that extrapolation of the $C \times t$ paradigm is relevant at lower concentrations in the region of the LOAEL of 0.7 mg/m³.

For the pulmonary region, a regional gas deposition ratio (RGDR) of 2.25 for an adult female rat was used for a rat-to-human extrapolation based on the U.S. EPA Human Equivalent Concentration (HEC) (OEHHA, 2008), which adjusted the 1 hr exposure to 9.45 mg/m³.

A default LOAEL-to-NOAEL uncertainty factor (UF) of 6 was applied based on the transient increase in total protein in BALF after exposure ended. A LOAEL-to-NOAEL UF=6 is used in cases where the acute effect is considered mild rather than severe. Total protein was increased at the lowest concentration of 0.7 mg/m³ three hr post-exposure. At one day post-exposure, total protein in BALF had returned to control levels in rats exposed to 0.7 mg/m³, although total protein was still elevated, and essentially unchanged, in rats exposed to higher PMDI concentrations.

For potential differences between rats and humans, the interspecies toxicokinetic UF=2 is applied to account for residual toxicokinetic differences when using the HEC approach. A default interspecies toxicodynamic UF= $\sqrt{10}$ is applied to account for use of key studies employing non-primate species and the lack of data for toxicodynamic interspecies differences.

For the intraspecies toxicokinetic UF_{h-k}, the most sensitive effect occurs in the epithelial tissues of the pulmonary region of the lung where the relative pulmonary minute volume to surface area ratio is 3-fold greater in infants compared to adults (OEHHA, 2008). Therefore, the pulmonary effects are predicted to be greater in infants and children, necessitating a UF= $\sqrt{10}$ to account for intra-individual variation. The toxicogenomics data for diisocyanates show gene variants associated with increased sensitivity up to 10-fold greater in workers developing diisocyanate-induced asthma. However, these findings address long-term exposures resulting in diisocyanate-induced asthma and is applied to the 8-hour and chronic REL derivations below. An intraspecies toxicodynamic uncertainty factor, UF_{h-d}, is applied to address the toxicodynamic diversity in the human population, including sensitive populations. In the case of

asthmagens such as MDI, OEHHA applies a $UF_{h-d} = 10$ to protect children with asthma. The cumulative $UF=3600$ results in an acute REL of $2.6 \mu\text{g}/\text{m}^3$ (0.26 ppb).

A BMCL of $0.4014 \text{ mg}/\text{m}^3$ was calculated for total protein in BALF one day post-exposure. Total protein in rats exposed to $0.7 \text{ mg}/\text{m}^3$ had returned to control levels at one day post-exposure, although total protein was still elevated in rats exposed to higher PMDI concentrations. The continuous models available in the U.S. EPA BMD suite (U.S. EPA, 2013) could not model the data for 3 hours post-exposure, the most sensitive time point for this effect. Applying the same acute REL derivation procedure for the one day post-exposure findings, but using the $\text{BMCL}=0.4014 \text{ mg}/\text{m}^3$ as the POD, and omitting the LOAEL-to-NOAEL $UF=6$, a comparison REL of $9 \mu\text{g}/\text{m}^3$ is calculated.

A comparison REL can be derived based on developmental toxicity reported in the developmental studies by Buschmann et al. (1996) and Gamer et al. (2000). The NOAEL and LOAEL for the Buschmann study was 3 and $9 \text{ mg}/\text{m}^3$, respectively. The NOAEL and LOAEL for the Gamer study was 4 and $12 \text{ mg}/\text{m}^3$, respectively. OEHHA concluded that running benchmark dose modeling on the data is not ideal due to all fetal effects below $12 \text{ mg}/\text{m}^3$ being within the historical control range for the strains, lack of a good dose-response curve, and control fetal incidences for effects that were generally lower than expected. Using a NOAEL of $3 \text{ mg}/\text{m}^3$ as a POD, no time adjustment is made and the 6-hr exposure is treated as one hour. Since the effects are systemic in nature, the default RGDR for the human equivalent concentration (HEC) adjustment is 1. To accommodate possible differences between rats and humans, the interspecies toxicokinetic and toxicodynamic UFs are assigned 2 and $\sqrt{10}$, respectively. The intraspecies toxicokinetic and toxicodynamic UFs are assigned 10 and $\sqrt{10}$, respectively, to account for intra-individual variation when using a sensitive animal model. Since the study examined a highly sensitive life-stage, fetuses, higher intraspecies UFs are not required. The cumulative UF is thus 200 and the comparison acute REL is $15 \mu\text{g}/\text{m}^3$. Thus, an acute REL based on pulmonary changes is adequately protective for developmental toxicity.

Evidence of cross-reactivity of diisocyanates suggests that a first exposure to MDI could result in diisocyanate-induced asthma in an individual previously exposed and sensitized to a different diisocyanate. As discussed above in Krone and associates, it is conceivable that infants could be sensitized with dermal exposure to diisocyanate-containing polyurethane products. Subsequent exposure to low-level airborne MDI could then result in asthmatic symptoms. However, definitive evidence that dermal sensitization results from exposure to these consumer products is lacking. Exposure to MDI at concentrations as low as 0.05 ppb ($0.51 \mu\text{g}/\text{m}^3$) have resulted in an asthmatic reaction in sensitized individuals (Suojalehto et al., 2011). This finding suggests that once primary sensitization occurs it is probably not possible to identify a no effect level to protect all individuals that acquired specific hypersensitivity to diisocyanates.

The same conclusion was presented in an International Consensus Report on Isocyanates (ICRI, 2002).

From the data available, the acute REL for MDI should prevent primary sensitization, and should be protective in many cases for previously sensitized individuals. The best animal evidence to date indicates that the dose that triggers an elicitation response in a rat asthma model is slightly below that causing acute pulmonary irritation in naïve rats (Pauluhn and Poole, 2011). This finding suggests that prevention of pulmonary irritation, on which the acute REL is based, should also prevent or reduce allergic sensitization.

8.2 MDI 8-hour Reference Exposure Level

<i>Study</i>	Reuzel et al., 1994a; Feron et al., 2001
<i>Study population</i>	Adult female Wistar rats (59 or 60/group)
<i>Exposure method</i>	Discontinuous whole-body inhalation exposure to 0-6.0 mg/m ³ PMDI
<i>Continuity</i>	6 hours per day, 5 days/week
<i>Duration</i>	104 weeks
<i>Critical effects</i>	Bronchiolo-alveolar hyperplasia
<i>LOAEL</i>	0.98 mg/m ³
<i>NOAEL</i>	0.19 mg/m ³
<i>BMCL₀₅</i>	0.118 mg/m ³ (0.0115 ppm)
<i>Time-adjusted exposure</i>	0.0421 mg/m ³ (0.118×6/24×5/7×20/10)
<i>Human equivalent concentration</i>	0.0947 mg/m ³ (RGDR 0.0421×2.25)
<i>LOAEL uncertainty factor</i>	Not applied
<i>Subchronic uncertainty factor</i>	1
<i>Interspecies uncertainty factor</i>	
<i>Toxicokinetic (UF_{a-k})</i>	2
<i>Toxicodynamic (UF_{a-d})</i>	√10
<i>Intraspecies uncertainty factor</i>	
<i>Toxicokinetic (UF_{h-k})</i>	10
<i>Toxicodynamic (UF_{h-d})</i>	10
<i>Cumulative uncertainty factor</i>	600
<i>Reference Exposure Level</i>	0.16 µg/m ³ (0.015 ppb)

Two chronic studies have been conducted in rats: one by Hoymann et al. (1998) in which rats were exposed to MDI for 18 hrs/day, 5 days/week, and a study by Reuzel et al. (1994a) in which rats were exposed to PMDI for 6 hrs/day, 5 days/week. Interstitial fibrosis in rats resulting from 18 hr/day exposure in the Hoymann study is the most sensitive endpoint among the two studies and strongly suggests a dose-dependent effect (see Table 4). However, for an 8-hr REL, it is more appropriate to consider the most sensitive endpoint from the

Reuzel study, since these rats were exposed closer to the time duration of the 8-hr REL (6 hrs/day, 5 days/week). The high incidence of fibrotic lesions at all dose levels in the 18 hr/day study were probably related to the longer daily exposures and the resulting reduction in recovery time between exposures (Feron et al., 2001; Pauluhn, 2011).

BMD modeling by OEHHA of the reanalyzed histopathology data in Feron et al. (2001) revealed that pulmonary bronchiolo-alveolar hyperplasia is the most sensitive endpoint (see Table 7) to use for REL derivation, with pulmonary fibrosis also considered a critical endpoint. The $BMCL_{05}$ for pulmonary fibrosis is about 2.5-fold higher than the $BMCL_{05}$ for bronchiolo-alveolar hyperplasia. Feron et al. (2001) suggested that the greater incidence of bronchiolo-alveolar hyperplasia when comparing the high dose groups from each study was due to the higher concentration used by Reuzel et al. (6.03 mg/m^3 vs. 2.05 mg/m^3 used by Hoymann et al.) resulting in a higher local tissue dose, and longer survival at the top dose in the Reuzel animals (average survival for the Reuzel study rats was 700 days, whereas average survival in the Hoymann rats was 518 days).

A time-adjusted exposure of 6 hrs/24 hrs x 5 days/7 days x 20 hrs/10 hrs was used for the 8-hr REL derivation, which accounts for extrapolation from the lab exposure paradigm to a continuous exposure and includes the assumption that half the daily volume of air intake in humans occurs during an active 8-hr period in accordance with our guidelines.

For pulmonary lesions, a regional gas deposition ratio (RGDR) of 2.25 was calculated based on U.S. EPA HEC methodology (OEHHA, 2008). This ratio represents an adult rat-to-adult human HEC adjustment. The RGDR was used, rather than the regional deposited dose ratio (RDDR), because both PMDI and monomeric MDI would be proportionally greater in the gaseous form at the level of the REL. Deposition sites for gaseous and aerosol forms of MDI may vary slightly within the respiratory tract. However, the hydrophobic nature of MDI would suggest primary deposition for both forms in the pulmonary region. For reactive chemicals such as MDI where lesions are formed in the pulmonary tract at points of tissue contact, toxicokinetic differences between rats and humans are not expected to be large and are partially accounted for with a HEC adjustment. Thus, an interspecies UF_{a-k} was assigned a 2, in accordance with our guidelines. An intraspecies toxicokinetic (UF_{h-k}) uncertainty factor of 10 was used to account the up to 3-fold greater pulmonary minute volume-to-surface area ratio in infants and children compared to adults, which is not accounted for in the rat-to-human interspecies HEC adjustment (OEHHA, 2008), and to account for differences in risk of diisocyanate-induced asthma, as observed in workers, based on genotype for a number of enzymes including GST, NAT, and epoxide hydrolase.

A default interspecies toxicodynamic UF of $\sqrt{10}$ was used. However, due to MDI's sensitizing potential and the greater susceptibility of children to the asthma-exacerbating effects of substances such as MDI (described in Section

5.2), an intraspecies toxicodynamic UF of 10 was applied. Dividing by a total UF of 600 gives an 8-hr REL of $0.15 \mu\text{g}/\text{m}^3$ (0.015 ppb).

Eight-hour and chronic RELs based on MDI sensitization or pulmonary function decrements could not be derived from any of the occupational studies presented in this summary due to lack of adequate dose-response data. However, some of these studies suggest that dermal exposure is a component in sensitization of workers handling MDI and exposed to airborne MDI in the workplace. Animal models support this theory, as sensitization via dermal exposure to MDI has been demonstrated with subsequent asthma-like symptoms resulting from inhalation exposure to MDI. However, since dermal exposure causing sensitization occurs in workers handling the material, it is not anticipated that the general population would be at risk from dermal exposure to MDI.

Very low levels of MDI exposure has caused asthmatic reactions in sensitized individuals, and it has been suggested that there is no measurable exposure level at which all sensitized individuals will be protected. Nevertheless, the 8-hour and chronic RELs should be protective for many individuals that could have been sensitized via dermal exposure or inhalation exposure from MDI and other diisocyanates.

8.3 MDI Chronic Reference Exposure Level

<i>Study</i>	Hoymann et al., 1998; Feron et al., 2001
<i>Study population</i>	Adult female Wistar rats (80/group)
<i>Exposure method</i>	Discontinuous whole-body inhalation exposure to $0\text{-}2.05 \text{ mg}/\text{m}^3$ MDI
<i>Continuity</i>	18 hours per day, 5 days/week
<i>Duration</i>	104 weeks
<i>Critical effects</i>	Pulmonary interstitial fibrosis
<i>LOAEL</i>	$0.23 \text{ mg}/\text{m}^3$
<i>NOAEL</i>	Not observed
<i>BMCL₀₅</i>	$0.0256 \text{ mg}/\text{m}^3$
<i>Time-adjusted exposure</i>	$0.0137 \text{ mg}/\text{m}^3$ ($0.0256 \times 18/24 \times 5/7$)
<i>Human equivalent concentration</i>	$0.0455 \text{ mg}/\text{m}^3$ ($0.0137 \times 3.32 \text{ RGDR}$)
<i>LOAEL uncertainty factor</i>	1
<i>Subchronic uncertainty factor</i>	1
<i>Interspecies uncertainty factor</i>	
<i>Toxicokinetic (UF_{a-k})</i>	2
<i>Toxicodynamic (UF_{a-d})</i>	$\sqrt{10}$
<i>Intraspecies uncertainty factor</i>	
<i>Toxicokinetic (UF_{h-k})</i>	10
<i>Toxicodynamic (UF_{h-d})</i>	10
<i>Cumulative uncertainty factor</i>	600
<i>Reference Exposure Level</i>	$0.076 \mu\text{g}/\text{m}^3$ (0.0074 ppb)

In the reanalysis of the Hoymann study by Feron et al. (2001), interstitial fibrosis was the most sensitive endpoint and exhibited a steep dose-response for this lesion. Seventy-nine percent of the animals in the low dose group (63 of 80 rats) of 0.23 mg/m^3 showed minimal to moderate grade fibrosis.

Application of the time adjustment (18/24 hrs x 5/7 days) and the RGDR (3.32) results in an adjusted POD of 0.0455 mg/m^3 . For reactive chemicals such as MDI where lesions are formed in the pulmonary tract at points of tissue contact, intra- and inter-species toxicokinetic differences are not expected to be large. Thus, following our guidance when applying a Human Equivalent Concentration adjustment, an interspecies UF_{a-k} was assigned a value of 2. We assigned a value of 10 to the intraspecies UF_{a-k} . The intraspecies UF_{a-k} accounts for the up to 3-fold greater pulmonary minute volume-to-surface area ratio in infants and children compared to adults (OEHHA, 2008), as well as differences in risk of diisocyanate-induced asthma, as observed in workers, based on genotype for a number of enzymes including GST, NAT, and epoxide hydrolase.

A default interspecies toxicodynamic UF of $\sqrt{10}$ was applied. The intraspecies toxicodynamic UF of 10 was used to address MDI's sensitizing potential and the greater susceptibility of children to the asthma-exacerbating effects of substances such as MDI described in Section 5.2. This gives a cumulative UF of 600, and chronic REL of $0.076 \text{ } \mu\text{g/m}^3$ (0.0074 ppb).

For comparison, the US EPA based its RfC (similar to a chronic REL) of $0.6 \text{ } \mu\text{g/m}^3$ on a benchmark dose analysis of the Reuzel study using basal cell hyperplasia of the olfactory epithelium as the critical effect (U. S. EPA, 1998a). Whereas the RfC was based on data for males only, our analysis utilized the data for female rats in the reexamination of the Hoymann data by Feron et al. (2001). OEHHA chose the interstitial fibrosis as the critical effect because this was the most sensitive endpoint for an exposure duration (18 hrs/ day, 5 days/week) that came closest to a continuous chronic exposure. In addition, OEHHA used a larger toxicodynamic UF_{h-d} than USEPA specifically to protect against the onset of asthma symptoms in children, and a toxicokinetic UF_{h-k} of 10 to account for differences in risk of diisocyanate-induced asthma, as observed in workers, based on genotype for a number of metabolic and protective enzymes.

8.4 MDI as a Toxic Air Contaminant Especially Affecting Infants and Children

Under Health and Safety code Section 39669.5, OEHHA establishes and maintains a list of Toxic Air Contaminants (TACs) that may disproportionately impact infants and children. OEHHA evaluates TACs for addition to this list as we develop Reference Exposure Levels for TACs. MDI was identified by the ARB as a toxic air contaminant (TAC) in accordance with section 39657(b) of the California Health and Safety Code (Title 17, California Code of Regulations, section 93001) (CCR, 2007). MDI has been shown to cause asthmatic reactions

in sensitized asthmatic adults in controlled exposure studies, and in non-sensitized children with asthma as well as asthma-like effects in normal children exposed acutely to the diisocyanate MDI in an accidental exposure (Jan et al., 2008). OEHHA considers asthma a disease that disproportionately impacts children, and thus chemicals that induce or exacerbate asthma are considered more impactful for children (OEHHA, 2001). In addition, an animal study has shown that younger rats are more sensitive to the acute effects of MDI than young adult rats (Reuzel et al., 1994b). In view of the potential of MDI to exacerbate asthma and the differential impacts of asthma on children including higher prevalence rates, OEHHA recommends that MDI be identified as a TAC that may disproportionately impact children pursuant to Health and Safety Code, Section 39669.5(c).

9.0 References

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